POTENTIAL OF HIGH-RESOLUTION SATELLITE IMAGERY FOR MAPPING DISTRIBUTION AND EVALUATING ECOLOGICAL CHARACTERISTICS OF TREE SPECIES AT THE ANGKOR MONUMENT, CAMBODIA

MIZUKI TOMITA¹, YOSHIHIKO HIRABUKI², YUJI ARAKI³, SHINJI TSUKAWAKI⁴,
BORA LY⁵ AND PEOU HANG⁵

¹Faculty of Informatics, Tokyo University of Information Sciences, Onaridai 4-1, Wakaba-ku, Chiba 265-8501, Japan; e-mail, tomita@rsch.tuis.ac.jp;

²Faculty of Liberal Arts, Tohoku Gakuin University, Tenjinzawa 2-1-1, Izumi-Ku, Sendai 981-3193, Japan;

³Faculty of Education, Saitama University, Shimo-Okubo, Sakura-ku, Saitama 338-8507, Japan;

⁴Institute of Nature and Environmental Technology, Kanazawa University, Kakuma-machi, Kanazawa, Ishikawa 920-1192, Japan;

⁵Authority for Protection and Management of Angkor and the Region of Siem Reap, Bangkoug Village, Ampil District, Siem Reap Town, Siem Reap Province, Cambodia

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ABSTRACT

Large trees play several vital roles in the Angkor monuments landscape. They protect biodiversity, enhance the tourism experience, and provide various ecosystem services to local residents. A clear understanding of forest composition and distribution of individual species, as well as timely monitoring of changes, is necessary for conservation of these trees. Using traditional field work, obtaining this sort of data is time-consuming and labour-intensive. This research investigates classification of very high resolution remote sensing data as a tool for efficient analyses. QuickBird satellite imagery was used to clarify the tree species community in and around Preah Khan temple, to elucidate differences in ecological traits among the three dominant species (*Dipterocarpus alatus*, *Lagerstroemia calyculata* and *Tetrameles nudiflora*), and to identify crowns of the dominant species.

Population structures of trees were determined using a 14.26ha study plot. Species name, DBH, height, height under the crown and crown area were recorded for all trees over 40 cm in DBH. Tree locations were also recorded so as to provide references for the imagery analysis. Ecological traits of the dominant species were estimated using regressions by an expanded allometric equation for both large and small trees, based on DBH, height, height under the crown and crown width.

The total number of species in the study plot was 45. From a spatial perspective, the three dominant species over 100 cm in DBH were segregated from each other. *D. alatus*, *L. calyculata* and *T. nudiflora* were concentrated, respectively, along the approach to the temple, near the centre of the complex, and on the walls of the monument. Object Based Image Analysis (OBIA) conducted using QuickBird satellite imagery, showed that crowns of *D. alatus* were largely determined by maximum of NIR layer and mean of digital number in

panchromatic layer. Differences in the parameters for both asymptotic height and spatial distribution among the dominant species, result from differences in ecological traits, and enhance the value of the tourism resource by providing a dramatic shift of forest scenery that can be enjoyed by visitors to the monument.

Key words: allometric relationships, object based image analysis, QuickBird imagery, spatial distribution, tree community

INTRODUCTION

Tropical and subtropical tree communities, particularly those including huge emergent trees, provide superb forest scenery for tourists, not only in old-growth forest, but also in forest remaining near rural areas. In the Angkor monuments region, huge tree species belonging to the families *Dipterocarpaceae*, *Lythraceae* and *Datiscaceae* provide spectacular scenery, and also help ameliorate physical degradation of the monuments. Recently, environmental problems such as illegal logging and atmospheric and groundwater pollution have increased, along with development of the tourism industry in Siem Reap. The huge remnant trees are thus important for conservation of both biodiversity and tourism resources. Protection of these trees requires an understanding of their ecological traits, and of their contributions in conserving the Angkor monuments landscape.

Field research on trees and forests is time-consuming, labour-intensive and requires special skills and training. In recent years, however, new technologies employing Geographic Information Systems (GIS) and remote sensing technology are being examined as tools for efficient identification of tree species, vegetation/tree mapping and classification of land-use (e.g. Alexander & Millington, 2000; Palace et al., 2008; Katoh et al., 2009; Hirata et al. 2009; Zhao et al., 2014). In particular, remote sensing using very high resolution data, such as QuickBird, offers the potential for streamlining the process of producing maps of individual trees. In the last three decades, advances in computer technology, earth observation sensors and GIS science has led to the development of “Object-based image analysis (OBIA)” as an alternative to the traditional pixel-based image analysis (Baatz & Schape, 2000; Blaschke & Strobl, 2001; Blaschke, 2010). OBIA works on objects which are produced by image segmentation. As an object is a group of pixels, characteristics such as the mean and standard deviation of spectral values can be calculated and used in the classification. Grouping pixels is an effective method for overcoming the so called ‘salt and pepper effect’ (Blaschke et al., 2000) that often plagues pixel based classifications; and many researchers now prefer OBIA to traditional pixel-based image analysis (Blaschke, 2010). This OBIA method, if able to extract canopies and identify tree species, would be an extremely useful and laborsaving tool for management of tree communities in and around the Angkor monuments region.

The objectives of the present study were: (1) to clarify the current status of the tree species and tree communities in the Preah Khan monumental complex, (2) to detect differences in allometric relationships between diameter and height and/or crown width, among the top three dominant species (*Dipterocarpus alatus*, *Lagerstroemia calyculata* and *Tetrameles nudiflora*), (3) to detect crowns of the dominant tree species in and around the temple using QuickBird satellite imagery, and (4) to evaluate the bands most effective for detecting crowns and identifying tree species. This study is still in the preliminary stages, particularly in regards to detection and classification of crowns and species.

METHODS

The Preah Khan temple, being one of the few monumental complexes at Angkor still surrounded by forests, was selected as the research site. Preah Khan is situated in the northern part of Angkor, Siem Reap, Cambodia (13°27'N, 103°52'E; ca. 56ha). The central sanctuary of Preah Khan was built in 1191 and prospered as a monastery and teaching complex. Like many of the other historical sites, Preah Khan was rediscovered in the late 19th Century. A protected forest surrounds the outer moat. This forest (ca. 150 ha) plays several important roles; (1) as a remnant of the Dipterocarpaceae-dominant seasonal-dry forest, (2) as a vital tourism resource and (3) as a developed ecosystem providing various services to the local people.

Population structures and spatial distributions of remnant trees were determined using a 14.26ha study plot. The scientific and local names, diameter at breast height (DBH), height (H), height under the crown, and crown area were recorded for all trees over 40 cm in DBH. This field research was conducted in March in 2008 and 2009, with the assistance of the Authority for Protection and Management of Angkor and the Region of Siem Reap (APSARA) staff. The number of other tree canopies (OTC) over each target tree was also recorded. An OTC value of zero indicates that the canopy of the target tree was exposed to the sky.

To provide references for detection of the dominant tree species using ArcGIS, all the tree location and crown data were overlaid onto a QuickBird satellite image (copyright, Digital Globe Inc.), which was acquired over the study area (during a period without clouds) in January 2004. The January 2004 image was selected because there was no suitable image available from 2005 to 2009. The QuickBird dataset consisted of single band panchromatic imagery (450 – 900 nm) with spatial resolution of 0.61 x 0.61 m, and four band multispectral imagery (blue:450 – 520 nm, green:520 – 600 nm, red:630 – 690 nm and NIR:760 – 900 nm) with spatial resolution of 2.4 x 2.4 m.

To describe differences in ecological traits of the top three dominant tree species, height (or crown width) and diameter relationships were regressed by a simple allometric equation (Eq. 1) and an expanded allometric equation (Eq. 2):

$$H = aDBH^b \quad \text{Eq. 1}$$

and

$$1/H = 1/aDBH^b + 1/c \quad \text{Eq. 2,}$$

where a (cm cm^{-b}) is the initial growth index, b (dimensionless) is the ratio of relative growth rate of H to that of DBH , and c (cm) is the asymptotic height (Ogawa et al. 1965, Yoda 1971, Kohyama and Hotta 1990). These parameters are the regression coefficients (Ogawa et al. 1965, Yoda 1971). Major axis regression was used to examine Eq. 1 (Sokal and Rohlf 1995). Eq. 2 has more parameters and fewer degrees of freedom (d.f.) compared to Eq. 1, so the sum of squares of residuals (RSS) of Eq. 2 is smaller than that of Eq. 1. Thus, nonlinearity of the allometric relationship on log–log plots was compared using analysis of variance; i.e., the significance of a decrease in the RSS was tested at the expense of reducing d.f. (Aiba and Kohyama 1997, Kimura and Simbolon 2002): $F = [(RSS1 - RSS2)/(d.f._1 - d.f._2)] / (RSS2/d.f._2)$, where RSS1 and RSS2 are RSS in the regressions of Eq. 1 and Eq. 2, respectively, and d.f.₁ and d.f.₂ are d.f. of the regressions of Eq. 1 ($n-2$) and Eq. 2 ($n-3$), respectively. Additionally, regressions for both large and small individuals, based on DBH, height, height under the crown and crown width, were calculated, utilising dominant trees below 40 cm in DBH outside the study plot.

The OBIA method was adopted for the land cover classification, and implemented using commercially available software, 'Definiens Developer (formerly known as eCognition)' (Zhou et al., 2008). The first step in OBIA classification is segmentation of the image into image object primitives. In 'eCognition' technology; a new image object level is created, or morphology of an existing object is altered (Definiens, 2008). Subsequent steps of classification and segmentation algorithms were then applied to perform a hierarchical classification. In the first step of making the process tree, pixels in which brightness was equal to zero were classified into "back ground". Next, with the exception of the "back ground" class, objects were classified as "shadow" by determining their threshold value in the intensity band using "Feature View", a tool that allows for visualising a feature (image band or ratio of bands) and determining threshold values for classification. All other image objects were grouped into the "not shadow" class by using a membership function with inverted similarity to the class "shadow". The inverted similarity is a rule describing all image objects that are "not in the class shadow". All objects in the "not shadow" class were classified into either "ruin" by determining the ruin threshold values in the saturation and hue bands; or "vegetation" (essentially "not ruin") by using a membership function with inverted similarity to the class "ruin". At this point, only "vegetation" remained to be classified. In this last step, the crown projection map exported from ArcGIS was used to classify the crown trees into individual species. In this step, because the satellite imagery reflected the surface of the canopy layer, we employed only with zero OTC value. Statistical characteristics such as mean and standard deviation of each band (blue, green, red, NIR and Panchromatic layer), minimum and maximum of NIR, Brightness and NDVI were extracted from each polygon of crowns, and used as the independent variable in the next analysis. Species name was also extracted and used as the dependent variable.

The text file, which included both of the statistical characteristics as independent variables and the species name as dependent variables for each crown tree, was imported as an R data frame (R Development Core Team 2008). Classification and regression tree (CART) analysis was then performed. In the terminology of tree-structured methods, the entire data set is referred to as the root node of the tree. This root node is partitioned into subsets of the data that then comprise subsequent nodes. If a node is not subject to further partitioning, that node is called a terminal node. The decision rules that define whether a particular observation in the data set belongs to a particular node are called splits. The process of tree construction involves determining, for each node, which of a myriad of possible splits best explains the variability in the dependent variable in terms of the independent variables, and then deciding whether a node should be terminal or should be subject to further splits. The measure of variability used in making this decision is the deviance, a generalisation of the familiar residual sum of squares. If the dependent variable is categorical or ordinal, the resulting tree is called a classification tree; if the dependent variable is continuous, the tree is a regression tree.

RESULTS AND DISCUSSION

The total number of species (DBH > 40 cm) in the study plot was 45. *D. alatus* showed the largest relative basal area (30.6%), followed by *L. calyculata* (27.3%), *Irvingia malayana* (6.2%) and *T. nudiflora* (5.7%) (Table 1). The frequency distribution for DBH for *D. alatus*, *L. calyculata* and *T. nudiflora* showed inverse J-shaped, unimodal and bimodal distributions (Fig. 1, 2).

Table 1: Summary of all trees observed in the study plot. DBH: diameter at breast height; BA: basal area; RBA: relative BA

Family	Scientific name	Max. DBH (cm)	BA (m ² /ha)	RBA (%)	No. Stems (plot)
Dipterocarpaceae	<i>Dipterocarpus alatus</i> A.DC.	201.0	6.24	30.59	84
Lythraceae	<i>Lagerstroemia calyculata</i> Kurz	230.0	5.57	27.31	85
Irvingiaceae	<i>Irvingia malayana</i> Oliver ex A.Benn.	136.5	1.26	6.18	29
Datiscaceae	<i>Tetrameles nudiflora</i> R.Br.	145.0	1.17	5.73	30
Ebenaceae	<i>Diospyros decandra</i> Lour.	141.4	0.89	4.39	19
Moraceae	<i>Artocarpus chama</i> Buch.-Ham.	124.2	0.56	2.72	16
Flacourtiaceae	<i>Hydnocarpus anthelminthica</i> Pierre & Gagnep.	120.5	0.51	2.49	22
Moraceae	<i>Ficus altissima</i> Blume	225.0	0.51	2.49	2
Sapindaceae	<i>Schleichera oleosa</i> Merr.	123.0	0.31	1.54	14
Alangiaceae	<i>Alangium salviifolium</i> (L.f.) Wangerin	81.8	0.27	1.35	15
Ebenaceae	<i>Diospyros curranii</i> Merr.	81.6	0.27	1.34	16
Moraceae	<i>Streblus asper</i> Lour.	65.9	0.20	0.99	15
Apocynaceae	<i>Astonia scholaris</i> (L.) R.Br.	127.1	0.20	0.98	7
Loganiaceae	<i>Strychnos nux-vomica</i> L.	82.4	0.19	0.94	8
Leguminosae-caesalpinioideae	<i>Dialium cochinchinense</i> Pierre	85.7	0.19	0.92	8
Moraceae	<i>Ficus callosa</i> Willd.	66.5	0.17	0.85	12
Moraceae	<i>Ficus callophylla</i> Blume	110.0	0.17	0.84	4
Sterculiaceae	<i>Pterospermum lanceifolium</i> Roxb. & DC.	96.3	0.16	0.81	7
Meliaceae	<i>Sandoricum koetjape</i> Merr.	93.5	0.15	0.73	9
Leguminosae-mimosoideae	<i>Samanea saman</i> Merr.	82.6	0.12	0.58	5
Sapindaceae	<i>Nephelium hypoleucum</i> Kurz	87.8	0.12	0.57	6
Sterculiaceae	<i>Pterocymbium tinctorium</i> Merr.	81.5	0.11	0.54	7
Dipterocarpaceae	<i>Anisoptera costata</i> Pierre	84.4	0.11	0.52	3
Lauraceae	<i>Beilschmiedia roxburghiana</i> Nees	71.3	0.10	0.50	6
Meliaceae	<i>Aglaiia edulis</i> Hiern	57.3	0.10	0.47	8
Moraceae	<i>Ficus drupacea</i> Thunb.	126.7	0.09	0.43	1
Combretaceae	<i>Terminalia bialata</i> Steud.	62.8	0.08	0.41	6
Polygalaceae	<i>Xanthophyllum glaucum</i> Wall.	67.5	0.07	0.33	4
Moraceae	<i>Ficus tinctoria</i> G.Forst.	67.8	0.06	0.31	3
Meliaceae	<i>Aphanamixis polystachya</i> (Wall.) R.Parker	60.4	0.06	0.31	5
Lauraceae	<i>Dehaasia</i> sp.	60.4	0.05	0.26	3
Chrysobalanaceae	<i>Parinari anamensis</i> Hance	65.8	0.04	0.20	2
Euphorbiaceae	<i>Mallotus philippinensis</i> Muell.Arg.	53.8	0.04	0.19	3
Dipterocarpaceae	<i>Hopea odorata</i> Roxb.	80.5	0.04	0.18	1
Euphorbiaceae	<i>Antidesma bunius</i> Spreng.	47.7	0.04	0.17	3
Leguminosae-caesalpinioideae	<i>Sindora siamensis</i> Teijsm. ex Miq.	74.0	0.03	0.15	1
Leguminosae-caesalpinioideae	<i>Senna siamea</i> (Lamarck) H.S.Irwin & Barneby	55.5	0.03	0.13	2
Meliaceae	<i>Walsura robusta</i> Roxb.	43.4	0.02	0.10	2
Leguminosae-caesalpinioideae	<i>Peltophorum dasyrachis</i> Kurz ex Baker	56.2	0.02	0.09	1
Myristicaceae	<i>Knema globularia</i> Warb.	52.9	0.02	0.08	1
Leguminosae-papilionoideae	<i>Dalbergia nigrescens</i> Kurz	52.4	0.02	0.07	1
Moraceae	<i>Ficus annulata</i> Blume	43.8	0.01	0.05	1
Ebenaceae	<i>Diospyros pilosanthera</i> Blanco	43.2	0.01	0.05	1
Rhizophoraceae	<i>Carallia brachiata</i> Merr.	40.3	0.01	0.04	1
Moraceae	<i>Ficus racemosa</i> L.	40.1	0.01	0.04	1
21 families	45 species	230.0	20.38	100.00	480

Fig. 1: Frequency distribution for all trees surveyed

Fig. 2: Frequency distribution for the top three dominant tree species

The inverse J-shaped distribution of *D. alatus* suggests that the species is regenerating continuously, most likely due to favourable conditions for establishment and survival of seedlings and juveniles. On the other hand, unimodal and bimodal distributions of *L. calyculata* and *T. nudiflora* suggest that regeneration of these species had occurred simultaneously. In this forest, therefore, it is considered that *D. alatus* is a late-successional species while *L. calyculata* and *T. nudiflora* are early- and/or mid- successional species.

Eq. 2 resulted in a better fit of the allometric relationships between *H* and *DBH* of *D. alatus*, *L. calyculata* and *T. nudiflora*, while those between *H* and crown width showed better fit in *D. alatus*, *L. calyculata* except for *T. nudiflora* (Table 2).

Table 2: Results of analysis of variance comparing simple allometric equation and expanded allometric equation in each species.

DH: diameter and height relationship; DCW: diameter and crown width relationship; n.s.: not significant

Spp.	n	Simple allometry		Expanded allometry			F
		a	b	a	b	c	
DH / <i>D. alatus</i>	197	0.32	0.64	0.91	1.06	59.52	157.75 ***
DH / <i>L. calyculata</i>	140	0.37	0.58	1.48	0.96	41.87	2.91 ***
DH / <i>T. nudiflora</i>	13	0.62	0.45	0.00	3.64	31.99	15.51 ***
DCW / <i>D. alatus</i>	197	-0.14	0.71	0.72	0.72	569043.56	4.70 ***
DCW / <i>L. calyculata</i>	140	-0.17	0.67	0.84	0.63	1019821.74	2.11 ***
DCW / <i>T. nudiflora</i>	13	-0.82	1.04	0.35	0.93	56.83	1.63 n.s.

With respect to the results of the regression between *DBH* and *H*, asymptotic height (i.e. parameter *c*) was greatest in *D. alatus* (59.5 m), followed by *L. calyculata* (41.9 m) and *T. nudiflora* (32.0 m). Parameter *b* for *D. alatus* was similar to that for *L. calyculata*, although both differed from that of *T. nudiflora* (Fig. 3).

Fig. 3: Allometric relationships for the top three dominant tree species

On the other hand, asymptotic crown width (i.e. parameter c) was difficult to estimate, although Eq. 2 showed a better fit than Eq. 1. When H_s and crown widths of stems that reached 150 cm in DBH were calculated using Eq. 2 with the estimated parameters for each species (Table 2), results of H were 45.0 m for *D. alatus*, 34.0 m for *L. calyculata*, and 32.0 m for *T. nudiflora*. In comparison, results of crown width were 26.6 m for *D. alatus*, 19.7 m for *L. calyculata*, and 37.0 m for *T. nudiflora*. These results suggest that stems in *D. alatus* are able to achieve greater height growth with the same DBH compared to stems of other species, while those in *T. nudiflora* are able to achieve greater growth in crown area. In the study site, large sized *D. alatus*, around 50 m in height, were distributed along the corridor leading to the temple, providing a spectacle subtropical forest landscape for visitors to enjoy. Visitors reaching the center of the monument on foot, on the other hand, are treated to a unique experience provided by the wide crown *T. nudiflora* with roots twisting around the temple walls. This dramatic change in scenery is made possible by species of tree that differ in their habitats, maximum attainable height and crown width.

From a spatial perspective, the top three dominant trees over 100 cm in DBH were segregated from each other. *D. alatus*, *L. calyculata* and *T. nudiflora* were concentrated respectively along the approach to the temple, near the centre of the complex, and on walls of the monument (Fig. 4). Negative spatial association-like distribution and differences in the parameters for asymptotic height resulted from differences in ecological traits among the tree dominant species, and is responsible for the dramatic shifts in scenery across the monument.

Fig. 4: Spatial distribution of stems for the top three dominant tree species

The number of species with OTC values greater than zero was 36, which accounts for 80% of the total 45 species observed in the study plot (Table 1, Fig. 5a). Mean and standard deviation of each band (blue, green, red, NIR and Panchromatic layer), minimum and maximum of NIR, Brightness and NDVI were extracted from polygons of OTC (Fig. 5b).

Fig. 5: Crown projection map (a) and results of object based image analysis (b).

Both maps overlaid onto QuickBird satellite imagery. OTC indicates the number of other tree canopies (OTC) over a target tree.

The best classification tree of the species in the extracted data set was shown in Figure 6. By means of CART analysis, the tree can be seen to have three terminal nodes (Fig. 6).

Out of 14 potential predictors in the model specification, only two were selected by the cross-validation method. The total number of polygons of *D. alatus* in root node was 82, of which those in the first partition and the second partition comprised 26 and 21 polygons, respectively (Fig. 6). 31.7% of polygons in *D. alatus* showed values greater than, or equal to, 852 in maximum of NIR layer, while 25.6% of those showed less than 181.3 in mean digital number of panchromatic layer. By using these values, 57.3% of the *D. alatus* polygons were detected in the study plot, although other tree species also composed a canopy layer of this forest. For labor-saving identification and mapping of not only *D. alatus* but also other canopy trees, more information such as flowering phenology, as well as a reconstruction of the OBIA process tree, is needed in further research.

Fig. 6: Result of decision tree using classification and regression tree (CART) analysis. Optimal pruning level was determined by using cross-validation method. Numerals separated by slash indicate frequency of polygons of each species. Frequency of polygons of *D. alatus* was represented by upper and fifth numeral from the left. Max and Mn5 indicates maximum of NIR layer and mean digital number of layer 5 (panchromatic layer), respectively. For more details, see Results and Discussion.

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