

The Cost-estimation of Mechanical Pre-treatment Lines of Municipal Solid Waste in Latvia

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Abstract. Production of refuse derived fuel from municipal solid waste in future shall play a strategic role in an integrated waste management system. The amount of landfilled biodegradable materials thus will be diminished according to provisions of the 1999 Waste Landfill Directive. The aim of this article is to evaluate cost effectiveness based on cost evaluation of the different complication of the waste pre-treatment equipment completion and based on regenerable waste quantities in Latvia. The comparison of cost estimates is done in 3 scenarios considering potential waste quantities in Latvia: Scenario I – planned annual waste quantity is 20 kT; Scenario II – 40 kT and Scenario III – 160 kT. An increase in amount of waste and processing capacity means the decrease in costs of mechanical pre-treatment of 1 ton of waste. Thus, costs of mechanical sorting line under different scenarios with capacities of 10 t h⁻¹, 20 t h⁻¹ and 80 t h⁻¹ are EUR 32 per t, EUR 24 per t and EUR 15 per t, respectively. Most feasible cost for a set of mechanical pre-treatment equipment for the capacity of 10 t h⁻¹ is EUR 32 per t by using rotating drum screener with the following manual sorting. Mechanical pre-treatment equipment of unsorted municipal waste is economically nonbeneficial, when the use of fine (biologically degradable) fraction is not possible. As the sorting of biodegradable kitchen waste is not developed under the current waste management system in Latvia, the lines for mechanical pre-treatment of household waste would be better to install in landfills.

Key words: mechanical pre-treatment, municipal solid waste, capital costs, operation costs.

Introduction

The National Waste Management Plan for 2013–2020 has set relatively high implementation objectives for the separate collection of organic waste, recycling and landfill in concordance with the requirements under the European Union (EU) Directive 1999/31/EC on the landfill of waste and the Directive 2008/98/EC on waste (European Parliament, 2008). It stipulates that only 35% of the amount of biodegradable waste (BDW) landfilled in 1995 can be landfilled in 2020. A problem for the national waste management is created by the requirement under the Directive 2008/98/EC that until 2020, as much as 50% of household waste must be recycled.

The European Commission's study (European Commission, 2012) on the implementation of waste management directives in the 27 EU member states indicated to the significant dependence of the Latvian waste management system on waste landfills, which, in its turn, has created a number of subsequent problems such as: a large quantity of landfilled waste; the reduction objectives for BDW landfilling have not been met, leading to a large percentage of BDW in the landfilled waste and a small amount of processed municipal solid waste. According to the Latvian

Environment, Geology and Meteorology Centre data, 77% of the 826 thousand t of municipal solid waste generated in 2010, and 58% of the 983 thousand t generated unsorted municipal solid waste in 2011 was landfilled.

Setting up sustainable landfills and creation of infrastructure conforming to environmental protection requirements are in the task list for Latvia upon joining the EU. In order to assure the fastest possible development of sustainable waste management, ten waste management regions (WMR) were defined for efficient planning of waste management and sustainable use of local resources and EU funds. Landfills of all WMRs are managed by inter-municipal companies, but the next step is to organise the waste management system which would assure both: the long and sustainable service of landfills and compliance with new requirements in the waste management. It must lead to prevention of waste, increasing of waste recycling and regeneration thus affecting volumes and type of landfilled waste.

Landfilling of untreated municipal solid waste is unable to meet the rising demands in the area of environmental protection. In 2010, the amount of emissions generated by the waste sector in Latvia

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reached 5.5% of the total greenhouse gas (GHG) emissions generated by the economic sectors (12 098 Gg CO₂ equivalent) (UNFCCC, 2012). The EU Directive 1999/31/EC on the landfill of waste stipulates that the pre-treatment of waste needs to be ensured in all the EU member states.

Regarding the waste management system in Latvia, a target has been set of decreasing emissions potentially causing the greenhouse effect from biodegradable waste in landfills. The National Waste Management Plan for 2013–2020 provided for the installation of mechanical biological treatment equipment in all waste management regions by the end of 2013 to divert waste to composting, incineration or other equipment where no waste sorting is established at source. The use of separated organic part as a fuel can be considered as one of the solutions to reduce the amount of the landfilled waste. Production of refuse derived fuel (RDF) from municipal solid waste in future shall play a strategic role in an integrated waste management system. Mechanical biological pre-treatment (MBP) is considered an alternative to waste incineration (Ritzkowski, Heerenklage, & Stegmann, 2006), although one of the products resulting from the MBP process is a fuel intended for burning. Therefore, mechanical biological pre-treatment of waste is widely used in Austria and Germany as of 2005 in order to decrease volumes of biologically degradable waste (Stegmann, 2005).

Mechanical biological treatment (MBT) is beneficial considering high cost of waste incineration and well developed system of collection of source sorted waste. Currently, there are no waste incineration plants in Latvia, but in the coming years options will be open for the refuse derived fuel (RDF) prepared from waste to be incinerated in neighbouring countries (Rimaitytė, Denafas, Martuzevicius, & Kavaliauskas, 2010; KUT, 2007). The RDF can be used as the high quality alternative fuel in cement production plant in Latvia.

Every municipality is organising management of waste itself according to binding rules of the municipality considering State and regional waste management plans. Introduction of sorted waste collection system is planned in each waste management region by 2013. According to requirements of EU directives and national laws, municipalities in cooperation with waste managers shall ensure that by January 1, 2015 each municipality provides all residents with sorted waste collection service, by collecting at least sorted glass, paper, metal and plastics.

Over the past 10 years, scientific research on alternative and renewable resources, including the use of waste for energy production has been

developing rapidly worldwide. In Latvia, research on opportunities for unsorted municipal solid waste treatment, including refuse derived fuel generation, which would facilitate the national economic benefits, thereby ensuring preservation and further improvement of the environmental quality is insufficient. One of the few studies (Pubule, Kamenders, Valtere, & Blumberga, 2014) has analysed seven biowaste management scenarios for Baltic countries by multi-criteria analysis, system dynamics modelling and a correlation-regression analysis. It is concluded there, that anaerobic digestion of separately collected biowaste would be the best solution for Latvia. Dāce *et al.* (Dāce, Bērziņa, & Bažbauers, 2010; Dāce, Pakere, & Blumberga, 2013) have developed a system dynamic model for analysing the dynamic behaviour of the integrated waste management system of the primary packaging.

The aim of this article is to evaluate cost effectiveness based on cost evaluation of the different complication of the waste pre-treatment equipment variations and based on regenerable waste quantities in Latvia. The tasks:

1. To detect the capacity for the waste pre-treatment based on amount of an unsorted waste in the regions.

2. To evaluate costs for the mechanical pre-treatment facilities of the municipal solid waste.

The article is a continuation of the study (Arina, Bendere, & Teibe, 2012; Arina & Orupe, 2012, 2013; Kalnacs, Arina, & Murashov, 2013; Teibe, Bendere, Perova, & Arina, 2012) of the assessment of unsorted municipal waste pre-treatment technologies in Latvia. The results presented in the article are important for development of the waste management system and regeneration in Latvia. Scientific novelty – the mechanical sorting lines for unsorted waste have been evaluated by mass flow analysis, equipment and its operation costs in Latvia, which had not been carried out in Latvia to date.

Materials and Methods

The mechanical sorting lines for unsorted waste have been evaluated by mass flow analysis and by tools of *Microsoft Excel*. The materials mass balance of waste separation was established by Equation 1 (Tchobanoglus, Theisen, & Vigil, 1993).

$$M_A = M_I - M_O + M_G \quad (1)$$

where M_A – rate of accumulation of material within the system boundary;

M_I – rate of flow of material into the system boundary;

M_O – rate of flow of material out the system boundary;

M_G – rate of generation of waste material within the system boundary.

Table 1

Number of residents in waste management regions in Latvia and number of municipalities split by number of residents as at 01.03.2011

| Regions of Waste Management | Population | | | Number of regions split by population | | | | |
|-----------------------------|------------|---------|---------|---------------------------------------|-------------|---------------|---------------|--------------|
| | Town | Rural | Total | Below 4000 | 4000-10 000 | 10 000-25 000 | 25 000-40 000 | Above 50 000 |
| Austrumlatgales | 48 356 | 45 321 | 93 677 | 2 | 2 | 1 | 2 | - |
| Dienvidlatgales | 106 979 | 80 017 | 186 996 | 2 | 4 | 3 | - | 1* |
| Liepājas | 105 590 | 50 543 | 156 133 | 5 | 5 | 1 | - | 1 |
| Malienas | 29 292 | 45 748 | 75 040 | 5 | 1 | 3 | - | - |
| Piejūras | 87 557 | 53 864 | 141 421 | 4 | 2 | - | 2 | 1 |
| Pierīgas | 757 889 | 135 075 | 892 964 | 2 | 11 | 6 | 1 | 1** |
| Ventspils | 51 587 | 21 721 | 73 308 | 1 | - | 2 | 1 | - |
| Vidusdaugavas | 59 932 | 49 598 | 109 530 | 7 | 6 | 2 | - | - |
| Zemgales | 81 975 | 92 290 | 174 265 | 2 | 4 | 3 | - | 1 |
| Ziemeļvidzemes | 49 964 | 117 073 | 167 037 | 9 | 9 | 4 | - | - |

*About 90 000 people are residing in Daugavpils city.

**About 650 000 people reside in the capital city Riga.

Source: Central Statistical Bureau of Latvia

Selection of capacity of mechanical pre-treatment facilities analysed

Firstly, the analysis of WMRs in Latvia was done. Analysis of cost estimates of mechanical sorting lines emphasises the number of residents in regions, which defines amount of waste to be disposed. The Table 1 provides data of density of residents in WMRs in Latvia and split of number of municipalities by number of residents for purposes of establishing necessary capacity of mechanical pre-treatment.

Pierīgas WMR has the highest number of municipalities, where the number of residents is from 4 000-10 000 and from 10 000-25 000 in each.

The amount of household waste on average is 50-70% from the total quantities of waste collected in the country. The Table 2 provides theoretically estimated amounts of household waste generated and actually disposed quantities of municipal solid waste by WMRs. The waste amounts were calculated, assuming that one resident produces 250 kg of waste per year.

The Table 2 shows that the majority of WMRs requires equipment with estimated waste processing capacity of 20 kT yr⁻¹ (Austrumlatgales, Malienas, Ventspils, Vidusdaugavas) and within the range from 30 to 40 kT yr⁻¹ (Dienvidlatgales, Liepājas, Piejūras, Zemgales and Ziemeļvidzemes). Only in

Pierīgas region a larger capacity of waste processing can be estimated (about 200 kT yr⁻¹). Theoretically mechanical pre-processing cost estimate scenarios by processing capacity have been determined considering theoretical estimates of household waste quantities generated. Of course, actual numbers can be different (lower), since sorted waste collection is planned in the country. Since the legislative acts do not set sorted collection of biodegradable waste from residents as a mandatory requirement, also the collection of other sorted waste (paper, plastics, glass and metal) is based on voluntary actions. In forthcoming five years no considerable decrease of quantities of waste disposed in landfills can be expected. Higher quantity of solid waste to be disposed in landfill along with household waste can be attributable to small businesses; however, it is possible that corporate entities will be those from whom sorted waste collection will be facilitated. Therefore, in cost estimate scenarios no waste from small businesses is considered.

The scenarios of evaluation costs for mechanical pre-treatment facilities of the municipal solid waste

The comparison of cost estimates for facilities is carried out in 3 scenarios, considering the potential waste quantities in waste management regions in Latvia:

Table 2

Theoretically estimated amount of household waste generated and the amounts of municipal solid waste actually landfilled in AAR

| Regions of Waste management in Latvia | Theoretical generated household waste, t (250 kg per inhabitant per year) | Landfilled in 2012 (code200301) t y ⁻¹ |
|---------------------------------------|---|---|
| Austrumlatgales | 23 419 | 17 745 |
| Dienvidlatgales | 46 749 | 44 692 |
| Liepājas | 39 033 | 32 351 |
| Malienas | 18 760 | 8 251* |
| Piejūras | 35 355 | 25 922 |
| Pierīgas | 223 241 | 293 790 |
| Ventspils | 18 327 | 14 392 |
| Vidusdaugavas | 27 383 | 32 582 |
| Zemgales | 43 566 | 35 961 |
| Ziemeļvidzemes | 41 759 | 11 259** |
| Total | 517 593 | 522 149 |

* Data of 2011;

**As of 2011 Ziemeļvidzemes WMR is actively involved in mechanic pre-sorting of household waste (code: 200301) and preparation for regeneration; therefore, amounts of household waste (HW) disposed in 2012 has significantly decreased, e.g., when comparing the year 2011, when 16 462 t of HW were disposed with 2010, when 28 002 t of HW were disposed. Sorted waste has been collected from residents in the Ziemeļvidzemes WMR already for more than 10 years; therefore, the total quantity of waste to be disposed was lower than in other regions.

Source: Latvian Environment, Geology and Meteorology Centre

- Scenario I – planned annual waste quantities 20 kT; planned capacity of equipment 10 t h⁻¹;
- Scenario II – planned annual waste quantities 40 kT; planned capacity of equipment 20 t h⁻¹;
- Scenario III – planned annual waste quantities 160 kT; planned capacity of equipment 80 t h⁻¹.

sewerage, waste water purification systems; fence, gates; improvement. The costs include also design and construction field supervision expenses. The comparison does not include price of the land, infrastructure of access roads and the garage is also not planned.

Construction and operational cost data of mechanical pre-treatment

The following formulas are used for cost estimation (2), (3) and (4):

$$C_I = C_C + C_E + C_M \quad (2)$$

where C_I – investments (euro);
 C_C – capital expenses (construction works) (euro);
 C_E – cost of equipment (new) (euro);
 C_M – management or maintenance costs (euro).

The capital expenses are personnel quarters (area 24 m²) (building is provided with infrastructure); operational buildings (modules, 12 m²); sorting area; hangar for waste sorting line; weights with calculation system; electrical and water supply system, lighting,

$$C_{Man.} = D + M + W + C_{el.fuel} + O \quad (3)$$

where $C_{Man.}$ – management or maintenance costs (euro);
 D – depreciation of buildings and equipment (per year) (euro);
 M – maintenance cost of buildings and equipment (euro);
 W – wages including taxes (euro);
 $C_{el.fuel}$ – electrical energy and fuel costs (euro);
 O – other expenses (concerned with production 10%; administration costs 10%) (euro).

$$C = (C_{Man.} + L + T - I) / C_{equipm} \cdot H_{work} \quad (4)$$

where C – costs (euro t⁻¹);
 C_{equipm} – capacity of equipment (t h⁻¹);

L – costs for landfilling (*euro*);
 T – costs of transportation (*euro*);
 I – income from selling of recyclables;
 H_{work} – working hours per year.

It was assumed that the mechanical sorting process of unsorted municipal solid waste would produce RDF, recycled raw materials (glass, metal (Fe), aluminium (Al)), biodegradable waste that can be used for further treatment, and the rest of the waste to be disposed in the landfill (Fig.1). The income from the selling of materials have been defined using the publicly available waste material prices in Latvia: RDF material – 14 *euro* per t, non-ferrous metals – 1100 *euro* per t, metals – 180 *euro* per t, glass – 7 *euro* per t. Considering current prices for collection of waste at the landfill effective in Latvia (2013), it was assumed that fee for disposal of one ton of municipal solid waste is 28 *euro*. The transportation costs have been established for a delivery distance of up to 100 km based on the cost of 1 vehicle-kilometre – 1.16 *euro* – indicated in the Transport Development Guidelines for 2014–2020 in Latvia.

Management costs are calculated considering prices in Latvia in 2013. The estimated equipment service time is 10 years (Heyer, 2001), whereas that

of constructions – 30 years. It is assumed that the construction depreciation ratio is 3.33%, equipment depreciation ratio – 5%, energy price – 0.20 *euro* per kWh, 11 of fuel – 1.20 *euro*. The annual employee salaries have been estimated in three groups: a director – 16 000 *euro*, engineers – 12 000 *euro*, employees – 8 000 *euro*. The planned number of working hours is 2000 h per year.

The average equipment prices have been obtained collecting the offers provided by equipment producers or distributors in different projects and tenders in Latvia, as well as using available sources of literature (Tchobanoglous, Theisen, & Vigil, 1993; CalRecovery & PEER Consultants, 1993; Caputo & Pelagagge, 2002; Eunomia Research & Consulting, 2002; FCM, 2004; Tsilemou & Panagiotakopoulos, 2004, 2007). Table 3 shows the ranges of the values of the characteristic capacities and costs for the equipment analyzed in this article.

The equipment has been assembled considering machinery necessary for waste transportation lines, etc. Items included in costs of three scenarios are given below in Table 4.

To assess the costs for different mechanical pre-treatment (MP) equipment sets and estimate the amount of the resulting recovered waste material,

Table 3

Characteristics and costs of mechanical pre-treatment equipment and data of literature

| Position of costs | Data from literature (Tsilemou, 2007) | | Data for article (for planned capacities) | |
|-------------------------|---------------------------------------|--------------------------------|---|--------------------------------|
| | Capacity (t h ⁻¹) | Purchase Price (<i>euro</i>) | Capacity (t h ⁻¹) | Purchase Price (<i>euro</i>) |
| Shredders | 0.40 - 30.00 | 11 700 - 103 600 | 10 - 80 | 270 000 - 950 000 |
| Bag breakers | 3.00 - 35.20 | 48 900 - 157 500 | 10 | 150 000 |
| Screens | 15.00 - 191.20 | 35 300 - 218 600 | 10 - 80 | 160 000 - 1 200 000 |
| Magnetic separators | 4.30 - 40.00 | 7 300 - 54 300 | 10 - 80 | 60 000 - 200 000 |
| Eddy current separators | 1.30 - 35.00 | 29 300 - 108 600 | 10 - 80 | 120 000 - 240 000 |
| Manually sorting cabin | - | - | 10 - 80 | 120 000 – 180 000 |
| Ballistic separator | - | - | 10 - 80 | 220 000 - 750 000 |
| Post-shredder | - | - | 10 - 80 | 80 000 - 350 000 |
| Press, baler | 31 | 74 000 | 10 - 80 | 150 000 - 350 000 |
| Front-loader | - | - | - | 125 000 |
| Universal loader | - | - | - | 60 000 |
| Containers | - | 300 000 | - | 28 500 - 88 500 |

Table 4

Mechanical pre-processing equipment sets for three cost scenarios

| Position of costs | Scenario I (10 t h ⁻¹) | Scenario II (20 t h ⁻¹) | Scenario III (80 t h ⁻¹) |
|---|------------------------------------|-------------------------------------|--------------------------------------|
| Shredder | x | x | x |
| Screener | x (drum screener) | x (disc screener) | x (disc screener) |
| Magnetic separator | x | x | x |
| Eddy current separator | - | x | x |
| Manually sorting cabin | x (with 6 working place) | x (with 8 working place) | - |
| Ballistic separator | - | x | x |
| Post-shredder | x | x | x |
| Press, baler | x | x | x |
| Front-loader | x | x | x |
| Universal loader | x | x | x |
| Containers (30 m ³ and 1.1 m ³) | x | x | x |

Table 5

Option of mechanical pre-treatment equipment set at capacity of 10 t h⁻¹

| Pre-treatment equipment | Option 1 | Option 2 | Option 3 | Option 4 |
|----------------------------|----------|----------|----------|----------|
| Containers | x | x | x | x |
| Front-loader | x | x | x | x |
| Universal loader | x | x | x | x |
| Bag breakers/Openers | x | - | - | - |
| Shredder | - | x | x | x |
| Manually pre-sorting cabin | x | - | - | - |
| Disc screens | - | x | x | - |
| Rotation drum screens* | x | - | - | x |
| Magnets separator | x | x | x | x |
| Eddy current separators | - | x | x | - |
| Manually sorting cabin | x | - | x | x |
| Ballistic separator | - | x | x | - |
| Post-shredder | x | x | x | x |
| Baler, press | x | x | x | x |

Note. *The size of the sieve of drum screener is 60 × 60 mm.

Scenario I (the planned waste quantities – 20 kT yr⁻¹; capacity 10 t h⁻¹) with constant capital costs analyses the costs of four additional options (Table 5, Fig.1). Material fractions acquired as a result of operation of

sorting lines are estimated on the basis of previous studies by the first author regarding the content of waste after its mechanical pre-treatment (Arina & Orupe, 2012, 2013).

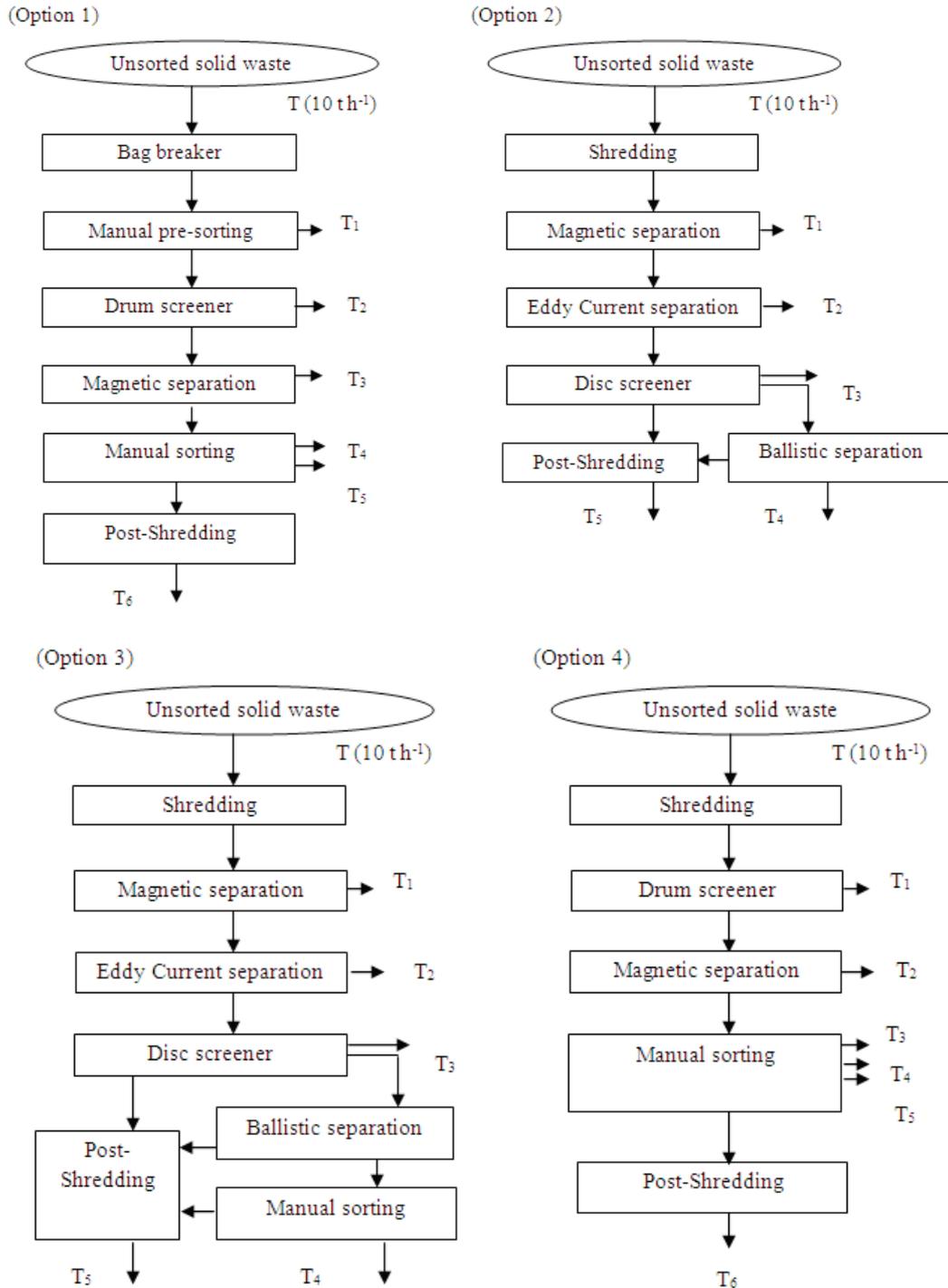


Fig.1. Flow diagrams of pre-treatment facilities for four options.

Note. Option 1: T_1 – Glass $\alpha = 0.03$; T_2 – Fine fraction to Biological treatment $\beta = 0.40$; T_3 – Ferrous Metals $\gamma = 0.02$; T_4 – Aluminium $\delta = 0.01$; T_5 – Residue to Landfill $\varepsilon = 0.09$; T_6 – RDF $\zeta = 0.45$;
 Option 2: T_1 – Ferrous Metals $\alpha = 0.02$; T_2 – Aluminium $\beta = 0.01$; T_3 – Fine fraction to Biological treatment $\gamma = 0.35$; T_4 – Residue to Landfill $\delta = 0.40$; T_5 – RDF $\varepsilon = 0.22$;
 Option 3: T_1 – Ferrous Metals $\alpha = 0.02$; T_2 – Aluminium $\beta = 0.01$; T_3 – Fine fraction to Biological treatment $\gamma = 0.35$; T_4 – Residue to Landfill $\delta = 0.16$; T_5 – RDF $\varepsilon = 0.46$;
 Option 4: T_1 – Fine fraction to Biological treatment $\alpha = 0.43$; T_2 – Ferrous Metals $\beta = 0.02$; T_3 – Glass $\gamma = 0.01$; T_4 – Aluminium $\delta = 0.01$; T_5 – Residue to Landfill $\varepsilon = 0.08$; T_6 – RDF $\zeta = 0.45$.

Option 1 – assumed that of recyclable materials (recovery ratios: glass $\alpha = 0.03$, ferrous metals $\gamma = 0.02$, aluminium $\delta = 0.01$) can be acquired, and the fine (incl. biodegradable) fraction (recovery ratio: $\beta = 0.40$) which, depending on the purity level, can be treated with aerobic and anaerobic methods and used for covering the landfill, and the course fraction (recovery ratio: $\zeta = 0.45$), which is planned to be prepared for regeneration, as well as waste to be disposed in the landfill (recovery ratio: $\varepsilon = 0.09$). Manual pre-sorting is intended to remove admixtures not suitable for further recycling of biological waste, especially the glass.

Option 2 – assumed that of recyclable materials (recovery ratios: ferrous metals $\alpha = 0.02$; aluminium $\beta = 0.01$) can be acquired, and the fine fraction (recovery ratio: $\gamma = 0.35$), the course (RDF) fraction (recovery ratio: $\varepsilon = 0.22$), as well as the medium fraction waste to be disposed in the landfill or incinerated (recovery ratio: $\delta = 0.40$).

Option 3 – assumed that of recyclable materials (recovery ratios: ferrous metals $\alpha = 0.02$; aluminium $\beta = 0.01$) can be acquired, and the fine fraction (recovery ratio: $\gamma = 0.35$), the course (RDF) fraction, and the part of manually sorted medium fraction waste, which is planned to be prepared for regeneration (recovery ratio: $\varepsilon = 0.46$ ($\varepsilon = 0.22+0.24$)), as well as medium fraction to be disposed in the landfill (recovery ratio: $\delta = 0.16$).

Option 4 – assumed that of recyclable materials (recovery ratios: ferrous metals $\beta = 0.02$; glass $\gamma = 0.01$; aluminium $\delta = 0.01$) can be acquired, and the fine fraction (recovery ratio: $\alpha = 0.43$), the course

fraction (recovery ratio: $\zeta = 0.45$), which is planned to be prepared for regeneration, as well as the waste to be disposed in the landfill (recovery ratio: $\varepsilon = 0.08$).

Paper and plastics recyclable materials can also be separated by applying mechanical sorting technologies; however, high quality materials can only be obtained from source separated paper and plastic waste. Therefore, in the estimates it is considered that these materials are intended for RDF production.

Results

The MP costs of municipal solid waste (MSW) in the three scenarios of mechanical sorting lines for MSW with the equipment capacities of 10 t h⁻¹, 20 t h⁻¹ and 80 t h⁻¹ are 32 *euro* per t, 24 *euro* per t and 15 *euro* per t, respectively. The equipment costs are 56% of the capital investments in Scenario I and 59% in Scenarios II and III (Fig. 2).

The economically most profitable solution is Scenario III, which – considering the potential waste amounts – can be implemented only in the Pierīga waste management region. With the increase in the cost of landfilling, Scenario I also becomes profitable, which – considering the available quantity – is suitable for a large part of waste management regions, assuming that separate waste collection will be developed.

Salaries of employees in Scenario I are 17%, in Scenario II – 15% and in Scenario III – 7% from management costs, while the electrical energy and fuel costs are 40%, 36% and 41%, respectively (Fig. 3).

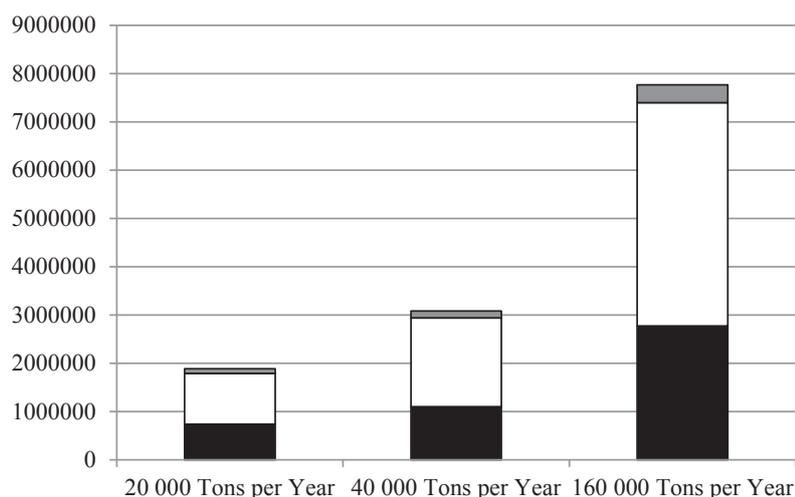


Fig.2. The comparison of investments (*euro*) for three scenarios of mechanical pre-treatment of municipal solid waste.

■ – Capital costs; □ – Costs of equipment; ▒ – Costs of projecting, organization of construction.

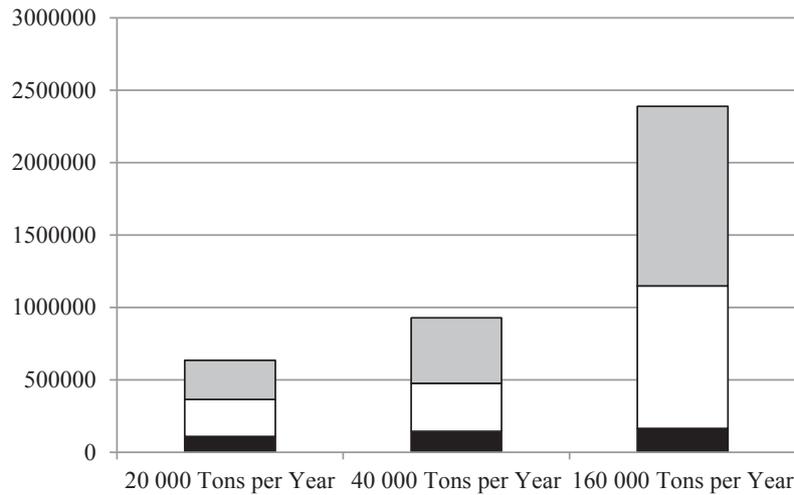


Fig.3. The comparison of overhead expenses (euro) for three scenarios of mechanical pre-treatment of municipal solid waste.
■ – Costs of personnel wage; □ – Costs of energy and fuel; ▒ – Other costs of management).

Table 6

Options of sets of mechanical pre-treatment technologies at capacity of 10 t h⁻¹ and costs for 1 ton of waste

| Costs | Option 1 | Option 2 | Option 3 | Option 4 |
|--|----------|----------|----------|----------|
| Costs, euro t ⁻¹ | 36 | 34 | 39 | 32 |
| Costs, incl. dispose in the landfill, euro t ⁻¹ | 38 | 45 | 43 | 34 |
| Costs, incl. income from realization and costs of transportation till 100 km, euro t ⁻¹ | 23 | 31 | 28 | 19 |
| Costs, incl. income from realization and costs of transportation till 250 km, euro t ⁻¹ | 32 | 41 | 35 | 26 |

Table 6 shows costs of four options mechanical pre-treatment technologies set versions for 1 ton of waste for equipment capacity of 10 t h⁻¹.

While the prices of equipment sets for Options 1 and 4 are equal, pre-treatment costs for 1 tonne of waste for Option 4 are lower, as it requires smaller labour force of manual work. The costs of Option 4 with a drum screener and manual sorting are the lowest; however, it must be taken into account that the proportion of the fine fraction (incl. biodegradable waste) after the use of a rotating drum screener is larger and contains a higher proportion of admixtures than after the disc screener.

For all four options, the income from selling the recovered materials exceeds the transportation costs up to the distance of 250 km.

The analysis of the results of the above options shows that in Option 1, as compared to Option 4, more material unsuitable for RDF generation can be sorted manually, thereby improving RDF quality although slightly increasing the amount of waste to be landfilled. While sorting lines in combination with manual sorting are more efficient in terms of costs, as higher quality material can be sorted, manual sorting of unsorted waste is considered to be work dangerous to health, which is why manual sorting is mainly used for sorting separated waste at source (Tchobanoglus, Theisen, & Vigil, 1993).

Discussion

Prices of equipment fluctuate and depend on equipment specification, quantity, geographic

location, transportation costs, discounts offered by sellers of machinery, etc. factors. It also has to be considered that every procurement tender contains certain provisions not allowing detailed price comparison, but rather just demonstrating overall kind of expenses.

According to Thiel (Thiel & Thomé-Kozmiensky, 2009), the waste disposal costs with the set of mechanical - biological waste treatment plants and waste incineration plants are equal in Germany – namely 100 *euro* per t. An average cost for disposal of waste in landfill in Latvia is 28 *euro* per t in prices of 2013, however, it is planned that the disposal price will be gradually increased.

According to data of the European Commission's research of 2003 (European Commission, 2003), the costs for RDF production from municipal solid waste in Belgium were 50-75 *euro* per ton.

Construction costs can be within the range of 40-70% from total investment, which in study scenarios, including designing are: in the Scenario I – 44%, in Scenarios II and III - 40% from total investment.

According to Heyer *et al.* (Heyer, Hupe, & Stegmann, 2001) referring to studies of Turk, Fricke, Hake the theoretical example of assessment of mechanical pre-treatment costs (prices of 1997) for a set of machinery intended for waste handling capacity of 30 kT yr⁻¹ depending on purposes of processing:

- Containers and pre-shredder – 24 *euro* per t;
- Containers, pre-shredder, magnetic separation, sieving (100 mm) – 27 *euro* per t;
- Containers, pre-shredder, magnetic separation, sieving (100 mm), post-shredder, sieving (40 mm) – 33 *euro* per t;

The example from literature sources has been compared against costs acquired within this study considering type of equipment. In comparison of expenses it must be considered that current prices of equipment are higher than those in 1997; however, studied prices in Latvia are considerably lower even without application of levelling coefficients. This could be explained by higher price for services in Germany.

According to Pubule *et al.* (Pubule, Kamenders, Valtere, & Blumberga, 2014) the minimal operational costs and capital costs for Baltic States are as follows:

- For anaerobic digestion of separately collected biowaste 28.00 *euro* and 376 *euro* per ton;
- For composting of separately collected biowaste 8.00 *euro* and 124.5 *euro* per ton of waste;
- For MBT with anaerobic digestion 28.00 *euro* and 372 *euro* per ton of waste;
- For MBT with composting 14.00 *euro* and 176 *euro* per ton of waste;

- For waste incineration with energy recovery 20.00 *euro* and 651 *euro* per ton of waste;
- For waste incineration without energy recovery 22.00 *euro* and 631 *euro* per ton of waste;
- For landfilling of biowaste 5.00 *euro* and 119 *euro* per ton of waste.

Data of estimation are not directly comparable as there are not costs for mechanical pre-treatment of the unsorted municipal solid waste in the literature.

When analysing economic benefits of mechanical biological pre-treatment of waste, unsorted household waste is basically divided into material groups: valuable materials, organic materials, light fractions, heavy fraction and residual materials. The purpose of mechanical pre-treatment is to separate valuable materials, which could be used and to separate currently not applicable materials, which shall be disposed in the landfill. According to descriptions of technology in the EU countries, available in sources of literature (Soyez & Plickert, 2001), when the purpose of technology is separation of fuel from waste, the fine fraction is planned to be further hygienised after the sieve with biological pre-treatment technologies and then once more sieved through finer material in order to separate sorts of waste suitable for fuel production. Thus, about 75-80% of materials can be reclaimed for production of fuel.

Sales of reclaimed materials such as aluminium in the future could decrease, when the planned deposit system for aluminium packaging will be introduced in Latvia (planned in 2013) (Brizga, Dimante, & Atstāja, 2012; Dāce, Bērziņa, & Bažbauers, 2010; Dāce, Pakere, & Blumberga, 2013).

According to data provided in sources of literature, quality compost can be obtained from source sorted biodegradable waste, not mechanically biologically pre-treated unsorted waste. Therefore, the current waste management system in Latvia, when source sorting of biodegradable catering waste is not developed, the lines for mechanical pre-treatment of unsorted municipal solid waste are better used in landfill.

Conclusions

1. Lower costs of mechanical pre-treatment of one ton of waste correspond to larger quantities of waste and processing capacity as shown by evaluation of the mechanical pre-treatment of the unsorted municipal solid waste. The costs of mechanical pre-treatment scenarios with the equipment capacities of 10 t h⁻¹, 20 t h⁻¹ and 80 t h⁻¹ are 32 *euro* per t; 23 *euro* per t and 15 *euro* per t, respectively.
2. The mechanical pre-treatment equipment for unsorted waste is economically little effective,

- if fine (biologically degradable) fraction cannot be used.
- The cost effective choice for mechanical pre-treatment facility of 10 t h⁻¹ is 32 euro per t, using rotating drum screener with the following manual sorting. The separation of the fine fraction from unsorted municipal solid waste can be ensured by a drum screener, which is about 33% cheaper than a disc screener.
 - Feasibility study shall be developed for each waste management region separately, since there are too many local factors affecting costs.

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