

Original paper

THE EXPOSURE OF HONEY BEES TO PESTICIDE RESIDUES IN THE HIVE ENVIRONMENT WITH REGARD TO WINTER COLONY LOSSES

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Abstract

The present studies are the second part of the research project dedicated to finding the causes for increased winter mortality of honey bee colonies. The aim of this task was to investigate incidents of overwintered colonies' death with regard to the potential interrelation to the exposure to pesticides. The samples of winter stores of bee bread and sugar food (honey or syrup processed by bees), beeswax and bees collected from apiaries with low and high rates of winter colony mortality were searched for acaricides used to control *V. destructor* and plant protection pesticides. The presence of acaricides used in apiculture has been detected in the 51% beeswax samples. The most abundant acaricide was tau-fluvalinate. The stores of bee bread and sugar food had a similar frequency of plant protection pesticide occurrence, ranging between 50–60%, but the number of active substances and their concentrations were substantially lower in sugar food samples. The most prevalent pesticides in pollen were fungicides (carbendazim and boscalid) and insecticides (acetamiprid and thiacloprid). Only a few pesticides were found in the several dead honey bees. The level of pesticide contamination (frequency, concentration, toxicity) of hive products and bees originating from apiaries with both a high and low winter colony survival rates, was similar, which created a similar extent of risk. Although the multiple varroacides and pesticides were present in the hive environment we not found unequivocal links between their residues and high winter colony mortality.

Keywords: bee bread, beeswax, honey bees, honey/syrup stores, pesticide residues, winter honey bee colony losses

INTRODUCTION

The widespread use of pesticides in crop protection management is indicated as one of the greatest hazards to bee health (Johnson et al., 2010; Goulson, 2013). Exposure to pesticides has received significant attention on account of the depopulation of honey bee colonies (*Apis mellifera* L.) persisting for several last years in many regions of the world (van der Zee et al., 2012). Honey bees are directly exposed to the plant protection products at the time of spraying, through pollen, and nectar collection and storage in the hive. The surface water and

guttation water produced by plants are also regarded as potential sources of pesticides (Johnson, 2015). Another cause of the pesticide presence in hives are acaricides commonly used in apiculture to control the parasitic mite *Varroa destructor*. These compounds are the main source of contamination of beeswax (Wallner, 1999; Mullin et al., 2010). The toxic effect of pesticides depends on their chemical properties, dose, route, time of exposure as well as the state of the bees (Smirle & Winston, 1987; Wahl & Ulm, 1983).

The most hazardous pesticide to bees are insecticides compounds (I), especially those of

the neonicotinoid, pyrethrin and organophosphate classes. Herbicides (H) and fungicides (F), which are not dedicated to combating insects, have a generally low acute toxicity to adult bees (Johnson, 2015). After insecticides, fungicides are the second group of pesticides that are most detected in honey bee colonies (Mullin et al., 2010), although their toxic effect is more frequently observed in bee larvae than in adults (Zhu et al., 2014).

A pesticide dose that kills 50% of exposed bees in a short time (24, 48 or 96 h) is the measure of their acute toxicity (LD_{50}). Sublethal doses are also significant to the behavior, physiology and health of individual bees (Desneux et al., 2007). However, it is more difficult to prove the effects of sublethal amounts on honey bee colony (Berry et al., 2013). Chronic contact and dietary exposure to multi-residues in pollen, honey and beeswax at the same time significantly threatens honey bee colonies. Pesticide mixtures may entail a synergistic effect and unexpected increase in toxicity in comparison to the simple sum of individual active substance toxicity (Pilling & Jepson, 1993; Mullin et al., 2010).

The presented study has been carried out as a part of a four-year research project that started after the survival of overwintered honey bee colonies significantly declined in Poland. The purpose of this project was to define which factors are responsible for the occurrence of this phenomenon in national apiaries. In 2009-2012 approximately 500 apiaries with a rate from 0 to 100% of winter losses of honey bees were sampled for bees and hive products. Collected data in the part about the natural food resources, bee management techniques and epidemiological status of pathogens and parasites and correlation of these factors with colony mortality over the winter has been described in Pohorecka et al. (2014). Results of the evaluation of honey bee colonies exposed to an environment contaminated with plant protection pesticides (PPP) and apiculture acaricides in relation to colony mortality are presented in this paper.

MATERIAL AND METHODS

The presented research was carried out in cooperation with beekeepers who voluntarily took part in the project. The participants had come from all of Poland's sixteen voivodeships; most had observed an increased rate in colony loss during winter but some had not experienced this problem at all. Beekeepers were instructed about the type and quantity of diagnostic material to be collected and the methods of sampling. The bee hive environment was assessed in the same colonies in which pathogen and parasite loads were measured (Pohorecka et al., 2014).

The majority of samples were collected in the early spring during the first inspection of colonies after the winter. In each apiary all types of samples were separately gathered from the same randomly selected dead colonies. If less than ten colonies died in the apiary, all sample types were taken from all colonies. If ten or more colonies died in the apiary, samples were taken from ten colonies. In the apiaries where all colonies survived the winter, up to ten samples were taken from living honey bee colonies. The following types of material were collected from each selected beehive: a minimum 150 ml of winter stores of sugar food (honey or syrup processed by bees) from brood combs (gathered into containers, in case whole combs with stores were sent, the sugar food was extracted in laboratory), piece of comb with a minimum 100 cells of bee bread, piece of empty brood comb (approx. 10x10 cm), and a minimum 300 dead of worker bees from the bottom board. In the laboratory, individual types of samples from each apiary were pooled and analyses were performed on bulk samples for each matrix. Beeswax samples were analyzed for the presence of acaricides used to control *V. destructor*. The bee bread, sugar food and honey bee samples were analyzed for pesticide residues from plant protection products. All the samples were stored at -20°C until analysis.

Acaricide analysis

Reagents and chemicals

Amitraz is an unstable molecule and degrades into three metabolites: *N*-(2,4-dimethylphenyl)-

N-methylformamidine (DPMF), 2,4-dimethylformanilide (DMF), and 2,4-dimethylaniline (DMA). DMF as the principal degradation product left in pure wax (Korta et al., 2001; Korta et al., 2003) was monitored for the presence of amitraz in the beeswax samples. Most of the referenced standard ingredients were purchased from commercial suppliers. Certified acaricide standards of DMF, coumaphos, tau-fluvalinate, bromopropylate, acrinathrine, bromfenvinfos were purchased from Dr. Ehrenstorfer (Augsburg, Germany). All working standard solutions were prepared by dissolving each compound in acetone at a concentration of 1000 mg/l and stored at -21°C. These reference materials were used for the preparation of beeswax samples fortified with known quantities of analyzed substances. Isooctane were supplied by SIGMA-ALDRICH Co. (St. Louis, MO, USA) and other chemicals – acetone, hexane – by Merck (Darmstadt, Germany).

Beeswax sample preparation

Analysis of acaricide residues was performed on bulk samples of beeswax formed after the melting of all sections of the combs collected from one apiary. This procedure was based on techniques developed by Jimenez et al. (2005) and Adamczyk et al. (2007). The pieces of combs were cut, placed into a 1L-glass jar and soaked for about 30 min in distilled water (about 500 ml) at 60°C. During the step of soaking, the mixture was vigorously stirred. Afterwards, the pieces of combs were placed in a colander and washed with about 1L of distilled water. Next, all the washed sections were placed on a gauze spread over a beaker with water and oven-melted at 80°C ($\pm 5^\circ\text{C}$) for 30 min. The mixture was cooled at room temperature so that the beeswax could solidify on the water. Purified wax samples were dried in the laboratory at a temperature of 20-25°C and stored at this temperature until analysis.

Beeswax foundation produced for organic beekeeping (Fachverband Bienenzuchtausrüstung, Gewähr für reines Bienenwachs 1566) was used for the method development and validation. The samples of beeswax foundation were spiked at concentration levels of 0.01,

0.05, 0.1, 0.5, 1.0, 5.0 and 10.0 mg/kg for each of the analyte. Beeswax foundation samples were weighed in glass jars/beakers and melted at 70°C using a thermostated mechanical shaker (Innova), after which acetone standard solutions of analytes were added. The mixtures were strongly mixed until the acetone was evaporated. The spiked beeswax samples were kept at room temperature until analysis, away from direct sunlight.

Extraction of analytes

The SPE technique developed by Adamczyk et al. (2007) was adopted for acaricide residue extraction from beeswax samples. A beeswax sample of 0.2 g was weighed in a polypropylene copolymer (PPCO) tube to which 10 μl of internal standard solution (bromfenvinfos 10 ng/ μl) and 10 ml of isooctane were added. The beeswax sample was completely dissolved at 70°C within 5 min using a Reacti-Therm III Heating/Stirring Module. The solution was transferred to the freezer for two hours, and then centrifuged for 15 min at -10°C and the liquid phase was decanted into a new tube. After that, 6 ml of isooctane was added to the sediment and the extracting procedure was repeated twice for each sample. Three extracts per sample were collected into the same tube and passed through Clearnet Florisil-SPE 1000 mg/6 ml Column (Agela Technologies) for acaricide residue extraction from the beeswax matrix. The eluate was evaporated to dryness in the Pierce Reacti-Vap™ III under nitrogen, then dissolved in 1 ml of isooctane and filtered through 0.2 μm PTFE syringe filter into an autosampler vial.

GC-ECD analysis

The GC-ECD procedure for acaricide residue analysis in beeswax samples was adopted from Fries et al. (1998) and Bogdanov et al. (2003). The analysis were performed with Gas Chromatograph with an Electron-Capture detector (ECD) (GC-14A, Shimadzu) and DB-35MS 30 m \times 0.25 mm \times 0.25 μm (Agilent J&W) chromatographic column. The oven temperature was programmed for 120°C ramped at 8°C/min to 250°C and held for 3 min, and then ramped again at 3°C/min to 310°C and held for 8 min. Injector and detector temperatures were 240°C and

Table 1

Limit of detection (LOD), limit of quantification (LOQ) and working range of acaricide analytical procedure in beeswax

Analyte	LOD (mg/kg)	LOQ (mg/kg)	Working range (mg/kg)	Linearity (correlation coefficient)	
				from – to	mean
Acrinathrin	0.10	0.50	0.50 – 10.00	0.9956 – 0.9972	0.9960
Bromopropylate	0.10	0.50	0.50 – 10.00	0.9974 – 0.9980	0.9973
Coumaphos	0.10	0.50	0.50 – 10.00	0.9958 – 0.9976	0.9969
DMF	0.01	0.05	0.05 – 5.00	0.9952 – 0.9986	0.9975
Flumethrin	0.50	1.00	1.00 – 10.00	0.9952 – 0.9972	0.9965
Tau-fluvalinate	0.10	0.50	0.50 – 10.00	0.9959 – 0.9984	0.9972

340°C, respectively. A split/splitless injector was used in the split mode with a split ratio of 1:11 and 1 µl of extract was injected on the column. The carrier gas was helium with a constant flow rate of 1 ml/min. The fortified beeswax samples were used for the system calibration with the following levels of each analyzed substance: 0.01, 0.05, 0.1, 0.5, 1.5 and 10 mg/kg. The internal standard method with bromfenvinfos was used for calculation of analyzed acaricide residues in the beeswax.

GC-MS analysis

DMFs were determined using the GC/MS technique on a Gas Chromatograph Mass Spectrometer (GCMS-QP 2010 Plus, Shimadzu) and ZB-5HT INFERNO 20 m × 0.18 mm × 0.18 µm chromatographic column (Phenomenex) (Lodesani et al., 2008). Separation and quantity determination of DMF were received for the following oven temperature program: 60°C for 1 min ramped at 20°C/min to 260°C and held for 5 min, and then ramped again at 30°C/min to 340°C and held for 30 min, injector temperature 250°C, and ion source and interface temperatures 250°C and 348°C, respectively. A split/splitless injector was used in the splitless mode (t = 1 min) and 1 µl of the extract was injected on the column. The carrier gas was helium with a constant flow of 0.43 ml/min. The DMF was identified out with full SCAN mode (50–650 Mz), and quantified with SIM (selected ion monitoring). The quantification and qualification mass ions for DMF were 120 and 149, respectively. The external standard method was utilised to quantify the fortified

samples of beeswax at concentrations of 0.01, 0.05, 0.1, 0.5, 1.0, 5.0, and 10.0 mg/kg.

Validation parameters

The validation parameters - detection and quantification limits, working range and linearity are summarized in Tab. 1. The quantification limit of DMF residues was calculated on a level of 0.05 mg/kg, 1 mg/kg for flumethrin, and 0.5 mg/kg for coumaphos, bromopropylate, acrinathrin, fluvalinate and deltamethrin. The ECD detector response of coumaphos, bromopropylate, acrinathrin, fluvalinate and deltamethrin residue determination in beeswax was linear in the concentration range of 0.5–10 mg/kg, and for flumethrin 1.0–10 mg/kg. The MS detector response of DMF was linear in the concentration range of 0.05–5 mg/kg.

The correlation coefficient of linear dependence of acaricide residue concentration and peak area in the working range reached a value of over 0.995 for all analyzed substances. The recovery of analyzed substances ranged from 60.2 (for DMF) to 96.0% (for coumaphos), 87.7% on average. The coefficient of variation of the repeatability and the within-laboratory reproducibility of the elaborated methods did not exceed 20% thus demonstrating the methods' sufficient accuracy and precision.

Pesticide analysis

Reagents and chemicals

The 80 pesticides analyzed in the multi-residue method are listed in Tab. 2. All pesticide standards had purity ≥ 93.3% (typically > 99%) and were purchased from Fluka, Sigma-Aldrich,

Table 2

A list of analyzed pesticides, method performance and limit of detection

Compound	RT (min)	Fragmentor voltage	MRM Transition (m/z)	Collision energy (eV)	LOD* (ng/g)		
					Nectar	Pollen	Bees
Acephate	0.87	60	184.0 → 143.0	4	5	10	5.0
Acetamiprid	4.60	80	223.1 → 126.0	20	0.1	0.2	0.3
Aldicarb	6.90	80	116.0 → 70.1	4	2	5	2.0
Aldicarb-sulfone	1.50	80	223.1 → 86.1	8	5	10	2.0
Aldicarb-sulfoxide	1.10	70	207.1 → 89.0	8	4	6	4.0
Azoxystrobin	13.00	120	404.1 → 372.1	10	0.4	1	0.4
Boscalid	13.21	130	343.0 → 271.1	35	4	10	4.0
Bromuconazole	12.62	150	378.0 → 158.9	29	4	9	4.0
Captan	5.56	80	225.1 → 86.5	8	1	2	1
Carbaryl	9.77	80	202.1 → 145.1	4	1	2	1.0
Carbendazim	2.66	96	192.1 → 160.0	17	0.1	0.3	0.5
Carbofuran	9.23	80	222.1 → 123.0	21	0.2	0.5	0.2
Carbofuran-3-hydroxy	3.34	90	238.1 → 163.0	12	0.6	3	0.6
Carbofuran-3-keto	7.15	80	236.1 → 179.0	9	3	5	3.0
Chloridazon	3.53	145	222.0 → 77.0	35	0.2	3	0.2
Chlorotoluron	9.63	120	213.1 → 72.0	33	0.5	1	0.5
Clothianidin	3.07	104	250.1 → 132.0	16	0.5	1	2.0
Cymoxanil	5.72	50	199.0 → 128.0	4	1	3	1.0
Cyproconazole	12.25	120	292.1 → 70.0	15	0.8	2	0.8
Desmedipham	12.19	84	318.1 → 136.0	26	0.1	0.3	0.1
Diethofencarb	12.51	85	268.1 → 124.0	30	0.5	2	0.5
Dimethoate	4.08	70	230.1 → 199.0	4	0.5	1	0.5
Fenbuconazole	13.69	120	337.1 → 70.0	20	0.9	5	0.9
Fenpropimorph	13.19	130	304.3 → 146.9	30	0.1	0.4	0.1
Fenpyroximate	18.33	130	422.2 → 138.0	33	0.8	1	0.8
Flonicamid	2.40	90	230.1 → 174.0	16	5	8	5.0
Fluquinconazole	13.26	100	376.0 → 307.0	20	3	4	3.0
Flurochloridon	14.11	100	312.1 → 292.0	29	5	10	5.0
Flutriafol	10.10	120	302.1 → 70.0	15	0.5	1.5	0.5
Formetanate	0.87	120	222.1 → 165.1	12	1	2	1.0
Hexythiazox	18.00	128	353.1 → 168.1	25	0.7	3	0.7
Imazalil	10.28	130	297.0 → 158.9	22	0.4	1	0.4
Imidacloprid	3.63	80	256.1 → 175.0	21	0.2	0.8	0.5
Indoxacarb	16.55	104	528.1 → 203.0	45	2	4	2.0
Lenacil	8.84	70	235.1 → 153.0	13	0.5	1	0.5
Linuron	12.25	110	249.0 → 132.9	30	2	3	2.0
Malaoxon	9.28	90	315.1 → 99.0	25	0.2	0.5	0.2
Malathion	13.89	80	331.0 → 99.0	21	0.3	0.5	0.3
Metalaxyl	10.45	90	280.2 → 220.1	9	0.2	0.4	0.2
Metamitron	3.05	110	203.1 → 175.0	13	1	2	1.0
Metconazole	14.00	120	320.2 → 70.0	25	0.7	2	0.7
Methamidophos	0.86	90	142.0 → 94.0	9	1	5	1.0
Methiocarb	12.51	130	226.1 → 124.0	25	0.4	1	0.4
Methiocarb sulfoxide	2.73	80	242.1 → 185.0	10	0.2	0.5	0.2
Methiocarb-sulfone	5.91	70	258.1 → 122.0	16	5	9	5.0
Methomyl	1.84	70	163.1 → 88.0	8	0.4	1	0.4
Methoxyfenozide	13.77	90	369.2 → 91.0	45	0.1	0.2	0.1
Monocrotophos	1.44	60	224.0 → 127.0	12	2	3	2.0

Myclobutanil	12.82	115	289.1 → 125.0	37	5	10	5.0
Napropamide	13.47	108	272.1 → 129.1	14	0.5	1	0.5
Omethoate	0.89	90	214.0 → 125.0	20	1	1.5	1.0
Oxadixyl	8.07	80	279.1 → 219.1	9	0.3	1	0.3
Oxamyl	1.53	85	237.1 → 72.0	10	0.2	0.6	0.2
Paclobutrazol	11.98	120	294.1 → 70.1	20	1	2	1.0
Paraoxon (ethyl)	10.75	90	276.0 → 219.9	9	0.4	0.9	0.4
Paraoxon-methyl	7.99	130	248.0 → 201.9	18	1	2	1.0
Parathion	15.23	90	292.1 → 235.9	8	5	10	5.0
Pencycuron	15.99	128	329.1 → 125.0	25	0.1	0.3	0.1
Pendimethalin	17.91	80	282.1 → 212.0	9	0.9	3	0.9
Phenmedipham	12.18	84	301.1 → 136.0	18	2	4	2.0
Prochloraz	14.35	90	376.0 → 308.0	9	0.6	1	0.6
Propamocarb	1.13	100	189.2 → 102.0	13	1.1	2	1.1
Propaquizafop	17.02	120	444.1 → 100.1	14	0.5	1	0.5
Pyridaben	19.36	110	365.2 → 147.1	25	0.1	0.2	0.1
Quizalofop-ethyl	16.73	120	373.0 → 299.0	18	0.1	0.3	0.1
Rimsulfuron	9.29	124	432.1 → 182.0	21	4	10	4.0
Spinosyn A	14.15	140	732.5 → 142.1	35	0.4	1	0.4
Spinosyn D	15.00	120	746.5 → 142.1	15	1	2	1.0
Spirodiclofen	19.58	90	411.1 → 71.1	13	0.5	0.8	0.5
Spiroxamine	11.54	120	298.3 → 144.1	18	0.1	0.2	0.1
Tebuconazole	13.48	120	308.1 → 70.1	22	0.5	5	1.0
Tebufenpyrad	16.84	155	334.2 → 144.9	26	1	4	1.0
Thiabendazole	3.50	120	202.0 → 175.0	25	0.1	0.4	0.1
Thiacloprid	6.75	90	253.0 → 126.0	20	0.1	0.4	0.1
Thiametoxam	2.23	80	292.0 → 211.0	9	0.1	0.3	1.0
Thiodicarb	9.43	84	355.0 → 88.0	12	0.5	4	0.5
Thiophanate-methyl	8.86	100	343.1 → 151.0	20	0.6	2	0.6
Triadimefon	13.07	120	294.1 → 69.0	20	5	10	5.0
Triadimenol	11.98	90	296.1 → 70.1	16	4	10	4.0
Triticonazole	12.36	115	318.1 → 125.0	38	5	5	5.0
Zoxamide	15.44	116	336.0 → 159.0	45	0.5	1	0.5

*LOD – limit of detection

Promochem (Institute of Industrial Organic Chemistry), Dr Ehrenstorfer, Riedel-de Haen and Supelco. Ultra gradient HPLC were purchased from Carlo Erba. Anhydrous magnesium sulfate (grit and powder), sodium chloride, sodium citrate tribasic and sodium hydrogencitrate sesquihydrate were ordered from Sigma-Aldrich. Acetic acid (HOAc) (98% purity), n-heksan (99% purity) were obtained from Chempur. Formic acid (98% purity) was purchased from J.T. Baker, and ammonium formate from Fluka. Both SPE sorbents PSA bonded (primary secondary amine) and Discovery DSC-18 were obtained from Supelco. Triphenylphosphate (TPP) and Tri (2,3-dichloropropyl) phosphate (TDCPP) were purchased from Fluka.

Individual standard stock solutions at 1000 µg/ml

were prepared through the dissolving of standards in acetonitrile, methanol or acetone, depending on the solubility of a particular pesticide, and were stored in amber glass flasks at -18°C. A working standard pesticide mixture at a concentration of 10 µg/ml in MeCN (acidified with 0.1% HOAc c(v/v)) was prepared from the stock solutions and kept at 4°C before use. This mixture was appropriately diluted in MeCN for the preparation of matrix-matched calibration and fortification standards. TPP was prepared at 0.5 µg/ml in MeCN (acidified 0.1% HOAc (v/v)), and added to the final extracts, blanks and matrix-matched calibration. A solution of 50 µg/ml TDCPP in MeCN (surrogate standard) was added throughout the entire procedure and was used to control the extraction step.

Sugar food, bee bread and honey bee samples preparation

The sample preparation procedure based on a modified QuEChERS (Quick Easy Cheap Effective Rugged Safe) (Anastassiades et al., 2003; Wiest et al., 2011) has been described in detail in a study by Pohorecka et al. (2012). In brief, the matrix (sugar food 10 g, bee bread, honey bees 5 g each) was vortex with TDCPP, deionised water, acetonitrile and n-hexane. The mixture was then mixed and centrifuged with MgSO_4 , NaCl, sodium citrate tribasic, sodium hydrogen-citrate sesquihydrate. MeCN supernatant was mixed and centrifuged with MgSO_4 , PSA and C18. The obtained extract, after being mixed with water, acetonitrile and internal standard TPP, and then after being filtered, was analyzed by LC/MS/MS.

LC-MS analysis

For the LC analysis, an Agilent 1200 HPLC system with a binary pump was used. The analytes were separated on the C18 analytical column of 100 mm x 2.1 mm and 1.8 μm particle size (Agilent Zorbax Eclipse Plus). For the mass spectrometric analysis, an Agilent 6410 Triple-Quad LC/MS system was applied. The data recorded was processed with the Mass Hunter software. Both processes were conducted under previously described conditions (Pohorecka et al., 2012).

Validation parameters

The method was prepared according to the requirements of guideline SANCO/825/00 rev 8.1 from 16.11.2010: "Guidance document of pesticide residue analytical methods". Mean recovery efficiencies for all bee collected matrices ranged between 70–110% and the relative standard deviation was less than 20%, thus demonstrating the sufficient accuracy and precision of the method. The LODs were estimated from the injection of matrix-matched solutions at concentration levels corresponding to a signal-to-noise ratio of about 3 for the quantitation ion and presence of the confirmatory ion as well. The LOD levels for nectar and pollen samples are included in Tab. 2. During validation, linearity of the method was evaluated in matrix-matched standards for each matrix. The calibration were found to be linear with correlation

coefficients greater than 0.99 for pesticides included in the method.

Statistical analyses

All the statistical analyses were carried out using Statistica 10 StatSoft®. The values obtained for the investigated parameters were compared between the group of apiaries with high losses (>10%) of bee colonies and group of apiaries with low losses ($\leq 10\%$) of bee colonies. The relationships between the investigated qualitative characteristics were evaluated with the Chi-square test. P-values ≤ 0.05 were considered significant. Due to a lack of a normal distribution of variables the Mann-Whitney U test was used for comparison of pesticide concentrations.

RESULTS

A total of 727 samples of honey bees, food stores and beeswax were analyzed for pesticide residues. Various compounds were found to be present in 44.3% of samples. The prevalence of pesticides in bee bread, sugar food and beeswax was similar ranging from 49.6 to 60.2% (Fig. 1), while in honey bees this did not exceed 13.5%.

Acaricide residues in beeswax

Acaricides used in apiculture were detected in 156 of 306 beeswax samples analyzed. The most frequent residue was tau-fluvalinate found in 39.9% of samples (Tab. 3). The positive samples contained on average 1.6 mg/kg of this compound, wherein a mean concentration of tau-fluvalinate has been exceeded in one third of samples. Coumaphos residues were found in the second highest frequency (14.7% of polluted samples) with a mean concentration of about 1.0 mg/kg. Only five samples contained coumaphos in quantities higher than the average. Detection of other pesticides ranged from 1.0 % (acrinathrin) to 3.3% (amitraz as DMF). All samples were found to have no flumethrin residues.

There were no significant differences in the level of beeswax contamination recorded between the apiaries with high and low winter colony losses. Both frequency of detection (Fig. 2) and amounts of all acaricides searched were approximated. Mean concentrations of tau-fluvalinate were 1.59 (sd ± 0.53) and 1.61 mg/kg

Table 3

Acaricide residues in beeswax samples (n = 306)

Acaricide	Number of positive samples	Proportion of positive samples (%)	Concentration (mg/kg)		LOQ*
			Mean	Max.	
Acrinathrin	3	1.0	5.14	10.97	0.5
Bromopropylate	8	2.6	0.99	2.28	0.5
Coumaphos	45	14.7	0.99	11.02	0.5
DMF (amitraz metabolite)	10	3.3	0.18	0.27	0.05
Flumethrin	nd	nd	nd	nd	1.0
Tau-fluvalinate	122	39.9	1.60	8.10	0.5

*LOQ - limit of quantification

nd = not detected

(sd \pm 1.47) in beeswax from apiaries with low and high rates of colony mortality, respectively (Mann-Whitney U test, $p = 0.138$). Coumaphos measured in the beeswax from these apiaries had mean concentrations of 0.71 (sd \pm 0.09) and 1.05 mg/kg (sd \pm 1.84), respectively (Mann-Whitney U test, $p = 0.342$). Multiple residues (more than one acaricide) were found in 7.0% and 8.1% of beeswax samples, respectively.

Plant protection pesticide residues in winter food stores and honey bees

A total of 123 samples of bee bread were analyzed, 60.2% of which were found to be contaminated. The most frequently detected residues were from fungicides (in 45.3% samples) followed in order from insecticides (32.0%) and herbicides (24.5%). In positive bee bread samples, 22 pesticides were noted: ten fungicides, eight insecticides and four herbicides (Tab. 4). At least

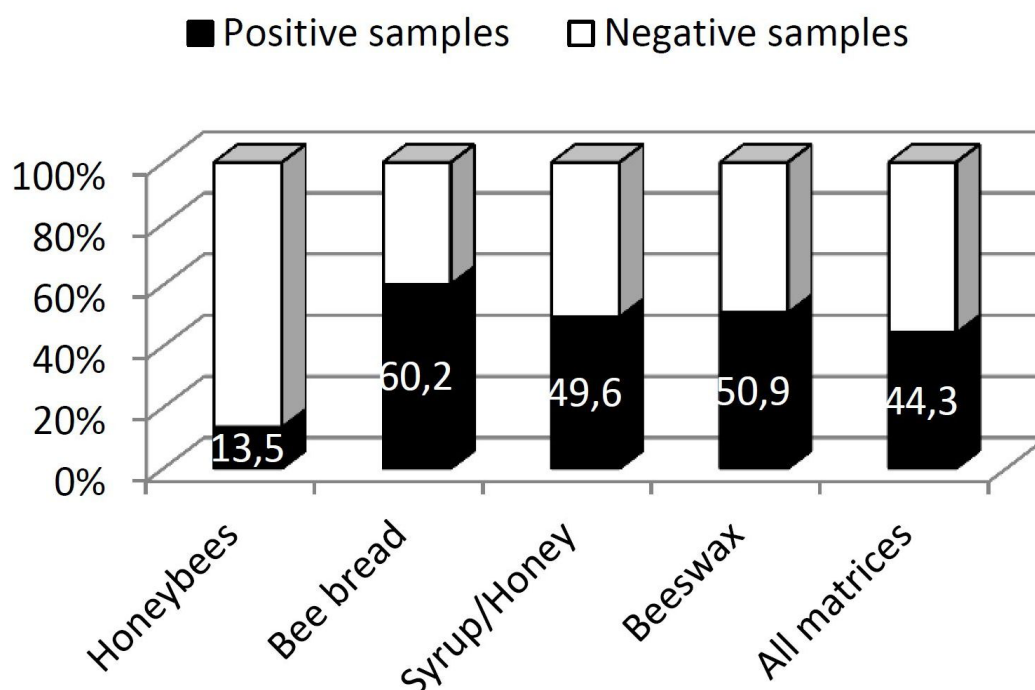


Fig. 1. Pesticide residues in bee-hive products and honey bees

Table 4

Pesticide residues in winter food stores

Compound	Bee bread (n = 123)			Syrup/Honey (n = 143)			LOQ* pollen/ honey
	Positive samples (%)	Concentrations (ng/g)		Positive samples (%)	Concentrations (ng/g)		
		Mean	Max.		Mean	Max.	
Acetamiprid (I)	15.4	10.3	32.8	31.5	1.6	19.5	1.0/0.5
Azoxystrobin (F)	7.3	6.9	12.3	nd	nd	nd	2.0/1.0
Boscalid (F)	18.7	124.9	1030.0	nd	nd	nd	12.0/5.0
Captan (F)	5.6	3426.7	9800.0	nd	nd	nd	2.0/1.0
Carbendazim (F)	30.1	7.1	44.6	12.6	1.8	3.1	1.0/0.5
Dimethoate (I)	9.7	36.4	162.7	nd	nd	nd	2.0/1.0
Fenpropimorph (F)	3.2	5.1	7.6	nd	nd	nd	1.0/0.5
Fenpyroximate (I)	0.8	3.2	3.2	nd	nd	nd	2.0/1.5
Imidacloprid (I)	nd	nd	nd	1.4	3.1	3.5	3.0/1.0
Linuron (H)	0.8	68.6	68.6	nd	nd	nd	3.0/4.0
Lenacil (H)	2.6	<LOQ	<LOQ	nd	nd	nd	2.0/1.0
Metamitron (H)	0.8	47.2	47.2	nd	nd	nd	3.0/2.0
Metalaxyl (F)	0.8	1.1	1.1	nd	nd	nd	1.0/0.5
Methoxyfenozide (I)	3.2	66.0	120.5	nd	nd	nd	0.5/0.5
Propamocarb (F)	0.8	4.4	4.4	nd	nd	nd	3.0/2.0
Prochloraz (F)	0.8	2.9	2.9	nd	nd	nd	2.0/1.0
Pendimethalin (H)	17.1	9.5	28.6	nd	nd	nd	4.0/2.0
Pyridaben (I)	0.8	3.2	3.2	nd	nd	nd	1.0/0.5
Thiametoxam (I)	0.8	4.3	4.3	0.7	13.3	13.3	1.5/0.5
Thiacloprid (I)	20.3	5.1	88.6	32.2	0.9	5.2	2.0/0.5
Tebuconazole (F)	9.7	99.7	387.0	1.4	1.2	1.3	6.0/1.0
Thiophanate-methyl (F)	9.7	34.7	110.6	nd	nd	nd	3.0/1.0

*LOQ – limit of quantification

F=fungicide, I=insecticide, H=herbicide

nd = not detected

Table 5

Pesticide residues in honeybees (n = 155)

Compound	Positive samples (%)	Concentration (ng/g)		LOQ*
		Mean	Max.	
Acetamiprid (I)	5.8	1.1	1.3	1.0
Carbendazim (F)	1.9	1.1	1.3	1.0
Dimethoate (I)	3.2	1.3	1.5	1.0
Fenpropimorph (F)	3.2	1.2	1.2	1.0
Fenpyroximate (A)	1.9	1.5	1.5	1.5
Imidacloprid (I)	1.9	4.1	5.3	2.0
Propamocarb (F)	1.9	<LOQ	<LOQ	2.0
Pyridaben (I)	1.9	1.4	1.8	1.0
Thiacloprid (I)	1.9	<LOQ	<LOQ	0.5

*LOQ – limit of quantification

F=fungicide, I=insecticide, H=herbicide

two pesticides were found in 38.2% of samples. Insecticides and fungicides co-occurred in 37.3% of positive samples. The most prevalent insecticides in pollen were thiacloprid (20.3%), acetamiprid (15.4%) and dimethoate (9.7%). The most frequent fungicide residues were carbendazim (30.1%) and boscalid (18.7%). The greatest amount of residues was estimated for captan in a quantity of 3,426.7 ng/g. The proportion of positive bee bread samples obtained from apiaries with high (54.0% positive samples) and low colony losses (66.0% positive samples) showed no significant differences in contamination (Chi-square test, $p = 0.704$).

Among 143 of sugar food investigated samples, 71 samples in total revealed the residues of four insecticides and only two fungicides (Tab. 4). The most abundant compounds were systemic insecticides thiacloprid (32.2%) and acetamiprid (31.5%). The concentration of all pesticide residues in syrup/honey samples was very low (ppb). The share of contaminated honey samples from apiaries differing in scale of lost

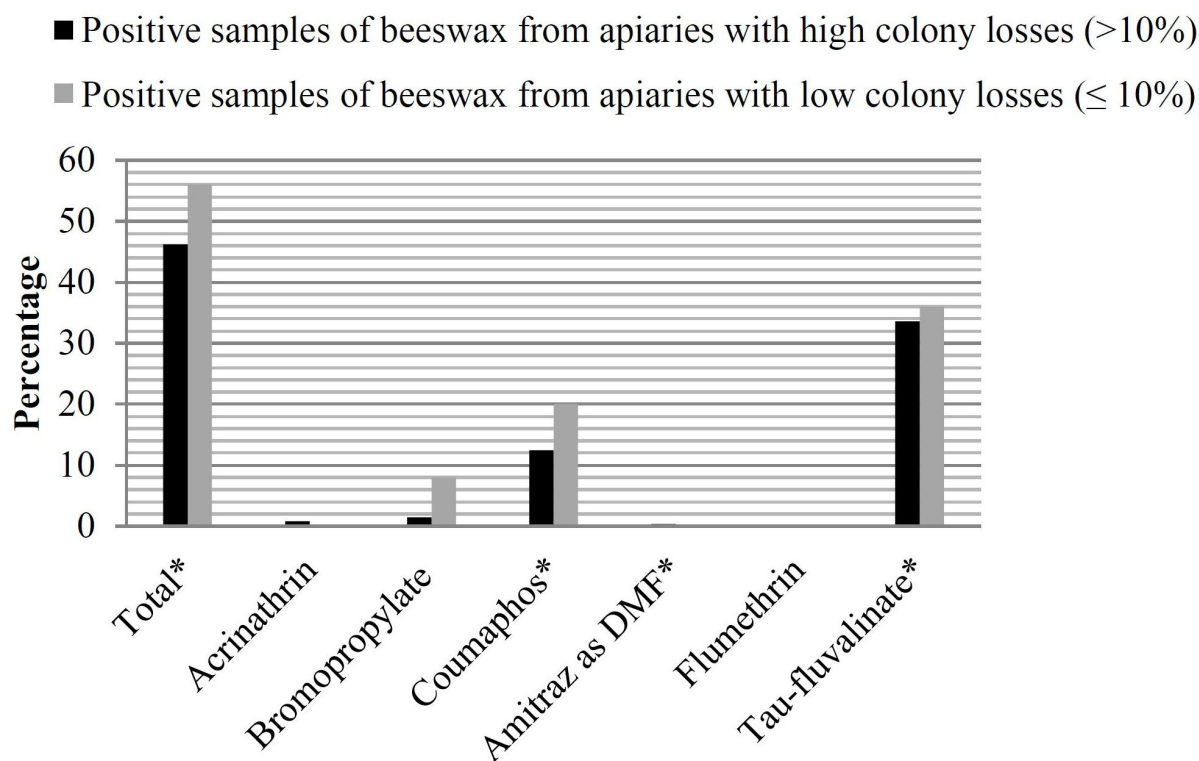
colonies overwinter was similar (Chi-square test, $p = 0.103$).

The results of the analysis of the 155 honey bee samples are presented in Tab. 5. The analysis of bees resulted in very rare detections. Small amounts of six insecticides and three fungicides were identified only in 21 samples. There were no significant differences in the proportion of bee positive samples between apiaries with high ($>10\%$) and low ($\leq 10\%$) loss rates of overwintering honey bee colonies (Chi-square test, $p = 0.363$).

DISCUSSION

Exposure to acaricide residues in beeswax

An overview of world data proved that acaricides applied to hives during treatments against Varroa mites accumulate within the hives, mainly in wax. However, the level of acaricide residue contamination in Polish beeswax had been unknown. Our study shows substantial varroa-cides load for this bee product, residues of at



*(Chi-square test, $p \leq 0.05$)

Fig. 2. Acaricide residues in the brood nest wax samples collected from apiaries with high and low colony losses

least one active ingredient have been detected in more than half of the analyzed samples. The main compounds that we found were tau-fluvalinate (39.9%) and coumaphos, although the latter was noticed in a much lesser proportion of samples (14.7%). According to the research of other authors, tau-fluvalinate and coumaphos belong to the most commonly found compounds (frequency of detection over 50%) in beeswax (Schanzes-Bayo & Goka, 2014). Tau-fluvalinate and coumaphos residues were present in 37% and 61% of beeswax samples collected from German apiaries (a large number of samples contains 1–5 mg/kg of both active ingredients) and in 61.9% (mean concentration 0.196 mg/kg) and 52.2% (on average 0.792 mg/kg) of French beeswax samples, respectively (Wallner, 1999; Chauzat & Faucon, 2007). In Italy, during the ten-year monitoring of beeswax present on the Italian market, 38% of samples were found to be contaminated with tau-fluvalinate (the average content in each year of the test ranged from 0.111 to 0.554 mg/kg) and 49% with coumaphos (average annual content 0.063–0.831 mg/kg) (Boi et al., 2016). Tau-fluvalinate (mean 1.310 mg/kg) was detected in almost all the commercial beeswax samples analyzed in Spain, while coumaphos was confirmed in only 3.7% of samples, at an average concentration 0.067 mg/kg (Serra-Bonvehí & Orantes-Bermejo, 2010). A definite broader beeswax contamination has been documented in USA, where both the frequency of fluvalinate and coumaphos detections, reached 98%, at a mean concentration of 7.3 and 3.4 mg/kg, respectively (Mullin et al., 2010). This means that most of the wax samples (83%) simultaneously contained residues of those pesticides.

The fact that the level of Polish wax contamination was comparable to what European authors were presenting, was surprising and also worrisome. Pesticides used in apiculture and agriculture are the main pollution source of chemical compound residues in beeswax (Johnson, 2015). In contrast to the above-mentioned countries, veterinary medicinal products with coumaphos and fluvalinate has not been authorized in Poland for the treatment of *V. destructor* in the

last several years. Moreover, the plant protection products containing coumaphos are not allowed for use in national agriculture, while tau-fluvalinate is approved only for oilseed rape protection under the Mavrik trade name. Nevertheless, the concentration of tau-fluvalinate and coumaphos residues in the beeswax turned out be the same or often even higher than those observed in France, Italy or Spain, where varroacides with tau-fluvalinate and/or coumaphos are approved. In such a context as this, the unauthorized varroacides would seem to be the main reason for Polish wax to be contaminated with coumaphos and presumably with tau-fluvalinate to large extent as well. This finding may be partly the result of the beeswax international trade.

Amitraz as an active ingredient of Apiwarol® and Biowar 500 (formulations permitted for apiculture) has been utilized the most and the longest by Polish beekeepers for chemical treatment against Varroa mite invasion (Pohorecka et al., 2014). Despite widespread application, amitraz (as metabolite DMF) has been detected in a very small proportion of samples (3.3%) and its mean concentration (0.18 g/kg) tenfold lower than tau-fluvalinate. The similar results have also been obtained in other European countries, where amitraz is registered for treatment against *V. destructor*. In Spanish and Italian beeswax the amitraz metabolites on average amounted 0.029 mg/kg (14% of positive samples) and 0.033–0.273 mg/kg (6% of positive samples), respectively (Serra-Bonvehí & Orantes-Bermejo, 2010; Boi et al., 2016). In accordance with 004/141/EC Commission Decision, in European Union amitraz is forbidden as active substance of plant protection products, which significantly limits of honey bee exposure to this chemical compounds. In the USA, where the Environmental Protection Agency (US EPA) authorizes the use of amitraz in agriculture and as a veterinary miticide (but not for its use in apiculture), amitraz metabolites (DMPF, DMA) were found in 60.0% (mean concentration 2.2 mg/kg) and 34% (mean concentration 0.7 mg/kg) of beeswax samples (Mullin et al., 2010).

Amitraz decomposes very rapidly in beeswax (Korta et al., 2001), so its application creates

Table 6

Relationship of the maximum residue detected in a sample of bee bread to the oral LD₅₀ for an adult worker honey bee, based on a consumption of 10 mg of pollen per nurse bee per day and during the 10-day nursing phase

Compound	Maximum residue (ng/g)	Maximum amount of pesticide residues ingested per nurse bee per day (µg)	Percentage of oral LD ₅₀ per nurse bee per day	Percentage of oral LD ₅₀ per bee during the 10-day nursing phase	Acute toxicity contact/oral** (LD ₅₀ µg/bee)
Acetamiprid (I)	32.8	0.00039	0.003	0.03	7.9/14
Azoxystrobin (F)	12.3	0.00015	0.0006	0.006	>200/>25
Boscalid (F)	1030.0	0.01236	0.007	0.07	>200/166
Captan (F)	9800.0	0.11760	0.1	1.0	25/91
Carbendazim (F)	44.6				>50/-
Dimethoate (I)	162.7	0.00195	1.1	11.0	0.12/0.17
Fenpropimorph (F)	7.6	0.00009	0.00009	0.0009	>100/>100
Fenpyroximate (I)	3.2				11/-
Imidacloprid (I)	nd				0.059/0.0037
Linuron (H)	68.6				
Lenacil (H)	<LOQ*				
Metamitron (H)	47.2				
Metalaxyl (F)	1.1	0.00001	0.000004	0.00004	141/269
Methoxyfenozide (I)	120.5	0.00144	0.001	0.01	>100/>100
Propamocarb (F)	4.4	0.00005	0.00005	0.0005	>100/99
Prochloraz (F)	2.9	0.00003	0.00005	0.0005	50/60
Pendimethalin (H)	28.6				
Pyridaben (I)	3.2	0.00004	0.007	0.07	0.053/0.55
Thiametoxam (I)	4.3	0.00005	1.0	10.0	0.025/0.005
Thiacloprid (I)	88.6	0.00106	0.006	0.06	36/17
Tebuconazole (F)	387.0	0.00464	0.005	0.05	>200/83
Thiophanate-methyl (F)	110.6	0.00132	0.001	0.01	>100/>100

* LOQ – limit of quantification

**Sources: Agritox database or US EPA database

F=fungicide, I=insecticide, H=herbicide

nd = not detected

a lower risk of wax contamination compared to the more stable lipophilic pyrethroid and organophosphorus compounds, although, on the other hand, the toxicity of amitraz breakdown products to honey bees has never been directly estimated (Johnson, 2015). Recent study showed that the most frequent pesticide residues identified in live honey bees were amitraz me-

tabolites (Kiljanek et al. 2017), so it could mean the honey bees may tolerate amitraz as an acaricide (Johnson et al., 2013). However, amitraz pre-treatment may increase the toxicity of tau-fluvalinate 5-fold, as a results of interactions between formamidines and pyrethroids. (Johnson et al., 2013).

Tau-fluvalinate is highly persistent in wax, so its

residues pose the highest risk by the long-term contact exposure of winter bees to contaminated combs. While most pyrethroids are highly toxic to honey bees, tau-fluvalinate is now considered as relatively less harmful and better tolerated in large part due to a rapid detoxification by cytochrome P450 monooxygenases (Johnson et al., 2006). Standard risk assessment complemented with new approaches that take into account time-cumulative effects over time conducted by (Sanchez-Bayo & Goka, 2014) confirms that the remaining tau-fluvalinate (and coumaphos also) at an average concentration level found so far (0.1 mg/kg) pose a very low risk through contact with pollen and with wax (as the authors suppose). Unfortunately, the average concentration of tau-fluvalinate and coumaphos in beeswax, 1.6 and 0.99 mg/kg, respectively, which has been documented in our monitoring, exceeds the average content of these substances in pollen at least tenfold.

In order to estimate the risk of bees being affected by contaminated beeswax we used a Hazard Quotients (HQ) expressed as a ratio of pesticide residues (in $\mu\text{g/kg}$) to their lethal dose (in $\mu\text{g/bee}$) (Stoner & Eitzer, 2013). Our calculations are based on the contact LD_{50} values 12 $\mu\text{g/bee}$ for tau-fluvalinate (PPDB) and 20 $\mu\text{g/bee}$ for coumaphos (PPDB), and the maximum concentration of these compounds in beeswax found in present study (the worst case scenario). Tau-fluvalinate and coumaphos had the HQ 675 and 551, respectively. In view of the European and Mediterranean Plant Protection Organization standards (EPPO, 2010) and European Food Safety Authority guidelines (EFSA, 2013) some researchers recognized pesticide load as "relevant" if had a HQ scores greater than 50 and "elevated" if the total HQ in a sample was 1,000 or more (Stoner et al., 2013; Kiljanek et al., 2017). Pesticide residues are substantially higher in beeswax, and transmission routes are less known in comparison with pollen or nectar. For this reason, Traynor et al. (2016) classified as "elevated" only samples of beeswax with a total HQ (the sum of all pesticide residues concentrations in ppb divided by their respective LD_{50} in $\mu\text{g/bee}$ for each residue in a given sample)

greater than 5,000. These authors found clear links between an increase in the total number of pesticides in wax and colony mortality. According to our results, the beeswax with tau-fluvalinate and coumaphos coming from hives with both high and low winter survival rates had similar contamination levels (frequency, concentration), which indicates a comparable extent of exposure to these pesticides. However, in the present study only acaricide residues were taken into consideration and it is highly likely that far more other pesticides (insecticides, fungicides) were present in beeswax.

Exposure to plant protection pesticide residues in winter food stores

In line with our expectations, the toxicological analysis of the winter carbohydrate (syrup/honey) and protein food stocks (bee bread) has shown a decidedly greater contamination of pollen. Although, the share of contaminated bee bread and sugar food samples was similar (60.2% and 49.6%, respectively), the number of active substances and their concentrations detected in sugar food were very limited and low. We noticed the presence of twenty-two pesticides in bee bread and only six in syrup/honey. These differences could be explained by the fact that most of food samples were obtained from overwintering honey bee colonies, so the majority of the carbohydrate food stocks had been formed from syrup. This conclusion is confirmed by our earlier findings, which showed a much higher contamination of real nectar/honey samples (Szczęsna et al., 2009; Pohorecka et al., 2012). Twenty nine compounds (among 30 surveyed) in 29% of honey samples collected in the summer from Polish apiaries were also found by Bargańska et al. (2013). The most frequently detected active substances were identified in the following order: profenofos (I), dimoxystrobin (F), diazinon (I). Eight of the thirteen pesticides were detected in Polish honey analyzed by Kujawski & Namieśnik (2011) and 87% of the samples were contaminated. In this study, the most common pesticide residue was clothianidin (I), which had been found in 65% of samples. Widespread honey contamination was also observed in Italy, where the majority of the

samples (94%) contained at least one of the pesticides tested (Panseri et al., 2014). Likewise, Chiesa et al. (2016) noted high residue frequency in organic honey produced in different geographic areas of Italy. Interestingly, pesticides have rarely been identified in French honey. No residues were detected in 57% of samples analyzed, the remaining contained a total 10 (from 39 investigated) of different analytes. The most frequent residues in honey were imidacloprid (21%) and its metabolite 6-chloronicotinic acid (18%) (Chauzat et al., 2011). Investigations of Austrian honey have shown acetamiprid, thiacloprid and thiamethoxam residues in samples (Tanner & Czerwenka, 2011). Residues in honey include mainly systemic compounds, among which the most commonly found are neonicotinoid insecticides, hydrophilic fungicides and herbicides (Schanzes-Bayo & Goka, 2014).

A total of 77 pesticides have been found in honey or nectar so far, whilst in pollen 124. The majority of PPP are fat soluble, so pollen is more susceptible to contamination. According to Schanzen-Bayo & Goka (2014) among the three factors relevant in risk assessment, which included quantity of residues, prevalence and toxicity, the latter which is the most important. Among the pesticides that we have found in pollen, fourteen belonged to fungicides and herbicides, which are generally less toxic to bees compared to insecticides, and only five (carbendazim, boskalid, pendimethalin, tebuconazole, and thiophanate-methyl) occurred more frequently (in the range of 10-30% of samples). Only three (thiametoxam, dimethoate and pyridaben) from ten insecticides present in pollen were classified as highly toxic but. The less toxic neonicotinoid insecticides – acetamiprid and thiacloprid – stood out due to their higher prevalence.

Exposure of winter bees to pesticide residues in pollen mainly occurs through the ingestion of contaminated pollen. We calculated the maximum daily doses of residues consumed with the maximum daily rate of bee bread (12 mg, Rortais et al., 2005) into a context of the hazard to a nurse bees (Tab. 6). Dimethoate and

thiametoxam residues posed higher risks, daily and ten-day intake equal to the consumption of 1% and 10% of their oral LD₅₀, respectively. According to EFSA guidance the environmental exposure of honey bees to pesticide residues higher than 1/10th of lethal dose would create a risk of initial bee acute toxicity. However, in our study, these two compounds appeared only once (thiametoxam) or in a few samples (dimethoate). Based on dietary risk evaluated for 77 compounds by Schanzen-Bayo & Goka (2014), we concluded that among the low residue loads of pesticide that we had found, the even more prevalent insecticides, thiacloprid and acetamiprid also posed a minor threat. Nevertheless, in more than half of the apiaries, the long-lived winter bees (consuming of 2 mg bee bread per day, Crailsheim et al., 1993) likely to have been exposed to sublethal pesticide doses up to 100 days. Continuous contact and dietary exposure to a low residues affect the behavior of individual bees and reproduction of honey bee colonies (El Hassani et al., 2008; Berry et al., 2013).

A risk assessments of oral and contact toxicity of the individual pesticides is insufficient in case of simultaneous exposure to different classes of pesticides, which may have a synergistic toxicity to bees. Iwasa et al. (2004) provided convincing evidence of how the addition of piperonyl butoxide and the fungicides trifluzole and propiconazole increased the acute toxicity (24-h LD₅₀, topical application) of acetamiprid and thiacloprid. Of the 123 analyzed pollen samples 22.0% were polluted with one compound and 38.2% were contaminated with more than one pesticide. Co-occurring pesticides did not constitute mixtures that had already shown synergistic toxicity (Schanzen-Bayo & Goka, 2014; Zhu et al., 2014), although only a very small fraction was studied for synergistic effects. In addition, the prevalence of pesticides in pollen from apiaries with high and low colony losses has been comparable. These effects suggest that the found levels of the pesticide contamination do not have a significant influence on the destruction of overwintering honey bee colonies.

Agrochemical contamination of bee bread not been evaluated in Poland before, so our results can be compared only to the data from other countries, although we are aware that there are large methodological and analytical differences among studies published so far e.g. different range of identified compounds, limit of detection (LOD), limit of quantification (LOQ) (Tette et al., 2016). Nonetheless, the diverse tests results may also be the result of various models of agrochemical treatments in particular regions of the world.

In some aspects our findings are similar to the results obtained in France and Germany where the positive samples of pollen reached 69% and 76% (pollen collected after the blooming period of oilseed rape) and were polluted with 22 and 42 active substances (32% with one pesticide) of the 41 and 258 searched, respectively (Chauzat et al., 2011, Genersch et al., 2010). In the French studies imidacloprid (33%) and 6-chloronicotinic acid (40%) were the most prevalent chemicals, while in the German ones boscalid (47%), coumaphos (37%), thiacloprid (33%) and terbutylazine (35%). In other parts of Europe less common pollen contamination has been shown. Pesticide residues were detected in 27% of Italian bee bread samples and 42% of Spanish pollen samples collected in spring, and in 31% of samples collected in autumn. The most frequently detected were fluvalinate and chlorfenvinphos (Bernal et al., 2010, Porrini et al., 2016). In the Netherlands, van der Zee et al. (2015) considered agricultural pesticides, neonicotinoids in particular, as one of many possible risk factors related to winter loss. From all the detected pesticides, only one group of neonicotinoids (thiacloprid, acetamiprid, and imidacloprid) was present in a sufficient number of samples. American researchers have noticed a definitely higher levels of pesticides in pollen, in which residues of 98 compounds (from 200 analyzed) were present in 99% of samples, with up to 31 different pesticides found in a single sample. A single sample contained on average 7.1 different pesticides; the most frequently found were residues of fluvalinate (88%) and coumaphos (75%), followed in order by chlorothalonil (53%),

pendimethalin (45%), chlorpyrifos (44%) and amitraz (31%). The insecticides chlorothalonil, aldicarb, carbaryl, chlorpyrifos and imidacloprid, fungicides boscalid, captan and myclobutanil, and herbicide pendimethalin were been assayed at ppm levels (Mullin et al., 2010). A direct relationships between pesticide residues found in stored pollen and colony losses was disclosed in none of the cited monitoring.

Plant protection pesticide residues in dead honey bees

We found nine pesticides in a total of 13.5% of honey bee samples, while in most of the samples only one compound was present. The quantity of all the pesticides we had detected in bees was much below the LD₅₀ values specified for adult honey bees (Tab. 6). Nevertheless, some of the identified substances characterized high toxicity to bees. These include dimethoate (LD₅₀ contact – 120 ng/bee, oral – 170 ng/bee, Sanchez-Bayo & Goka, 2014), pyridaben (contact LD₅₀ – 53 ng/bee, oral – 55 ng/bee, Sanchez-Bayo & Goka, 2014) and imidacloprid (contact LD₅₀ between 18-104 ng/bee, oral LD₅₀ 4-81 ng/bee, Blacqui re et al., 2012). However, these chemicals cannot be considered the cause of mass loss of colonies, because only a few samples were contaminated with them.

Obviously, some colonies were exposed to higher doses of pesticides as clearly indicated by results of the contamination level of wax and pollen stored in the nests of the surveyed bee colonies (Tab. 3 and 4). Pesticides can be biotransformed and/or excreted, which makes them impossible to detect. Furthermore, we analyzed the bees that had died in the winter and the time that had passed, since that could have also caused the breakdown of pesticides. Nonetheless, in samples of poisoned honey bees, there are observed a significantly higher frequency and quantity of pesticide residues in compared to samples of live honey bees (Kiljanek et al., 2017). Simultaneous analysis of 200 pesticides and metabolites conducted by these authors (2016) revealed 57 pesticides in poisoned honey bees. In a total of 74 samples of poisoned honey bees, only one sample was free of pesticides. The most abundant active

substances were highly toxic organophosphates: chlorpyrifos (38 samples), dimethoate (30 samples) and neonicotinoid clothianidin (22 samples). The concentration of several pesticides exceeded the LD_{50} in some samples. Severe contamination of honey bees from suspected pesticide poisoned colonies has been noted by Łozowicka (2013) and Bargańska et al. (2014). All samples were positive, and a large number of them contained more than one pesticide and with high residue concentration. The pesticides most often found in honey bees were highly toxic organophosphorus insecticides (chlorpyrifos heptenophos, methidathion) and pyrethroids (bifenthrin, cypermethrin). Similarly, analyses of samples received after the honey bee death incidents that occurred in Greece confirmed the presence of pesticide residues in 73% of them. In the majority of cases the bees were polluted with highly toxic pesticides: clothianidin, chlorpyrifos ethyl, thiamethoxam and imidacloprid (Kasiotis et al., 2014).

The researchers that used honey bees as bioindicators of environmental contamination with plant protection products found that even more extensively contaminated hive products (compared to a level detected in our experiments) did not necessarily cause the death of bee colonies. When the bee colonies were exposed to 22 pesticides present in pollen (70% positive samples), to 10 compounds in honey (43% samples), and to 15 pesticides in beeswax (65% samples), only the residues of 25 analytes were found in 44% samples of live honey bees (Chauzat et al., 2011). However, the majority of bee samples contained only one compound and the concentration of the most abundant substances (imidacloprid and 6-chloronicotinic acid) was very low. Along with the increased exposure (99% of pollen samples with 98 pesticides and 99% of beeswax samples contaminated with 87 compounds) the prevalence of pesticides in honey bees drastically increased – a total of 46 substances in 91% of contaminated bee samples (Mullin et al., 2010).

The broad analysis of the pesticide residues in the hive environment conducted with the aim of searching for the causes of the high mortality of

overwintered honey bee colonies has provided the first data on this issue in Poland. Acquired knowledge about exposure to pesticides (their prevalence and residue loads) of winter bees allowed the threat posed by these compounds to be assessed. We took into account various sources and routes for the exposure such as ingestion of contaminated food and contact with polluted beeswax as well as the acute, chronic and mixture of toxicities. We found that samples of beeswax were mainly contaminated with tau-fluvalinate, whose contact toxicity is recognized as low. Although, the bee bread contained the greatest number of pesticide residues, we assessed that the level of contamination created a relatively low toxicological hazard for colonies. Sugar food, which had been formed primarily from syrup, additionally reduced the potential source of pesticide residues. Similar exposure to pesticide residues in apiaries with high and low rates of winter colony mortality was an additional argument leading to the conclusion that in the surveyed apiaries contamination of the hive environment with pesticides was not the main reason for the decline of colonies during the winter. While, in apiaries with high colony losses the assessment of the epizootic state of these colonies showed significantly higher *V. destructor* infestation, the prevalence of deformed wing virus and acute bee paralysis virus (Pohorecka et al., 2011; Pohorecka et al., 2014). Finally, the present results confirm the previously proven thesis that in most national apiaries the mentioned above pathogens are a primary and direct cause of the increased winter mortality in honey bee colonies, but this does not mean that chronic exposure to pesticide does not have a side effect on bees. It would be expected that the co-occurrence of pesticide residues and pathogens/parasites may diminish the threshold of infestation/infection at which the colony collapses.

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