

HELMINTHOLOGIA, 47, 1: 58 – 66, 2010

Abundance and diversity of soil nematodes as influenced by different types of organic manure

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Summary

Nematode abundance and diversity from different types of organic manure soil treatments were investigated in a long-term field experiment carried out in Qu-Zhou experimental station, China Agricultural University. The composts used in the experiment were a traditional compost (C) (60 % straw, 30 % livestock dung, 5 % cottonseed-pressed trash and 5 % brans), traditional compost and chicken dung compost (60 % straw, 30 % chiken dung, 5 % cottonseed-pressed trash and 5 % brans) added with effective microorganisms, EMC and EMCDC respectively. Six treatments were arranged according to a randomized block design with three replicates per treatment. Treatments were incorporation into the soil of compost EMC, EMCDC, and C each at the rates of 7.5 and 15 t/ha. Plots were sown with winter wheat (*Triticum aestivum* L.) and summer maize (*Zea mays* L.) every year from 1997 to 2004. Overall, 28 nematode genera were found. Seven genera were bacterivores, 3 genera were fungivores, 13 genera were plant-parasites and 5 genera were omnivores-predators. The *Rhabditis*, *Cephalobus*, *Helicotylenchus* were dominant genera in the present study. The bacterivores and plant-parasitic nematodes were the dominant trophic groups except in C treatment. Nematode abundance per 100 g dried soil ranged from 372 to 553. Addition of effective microorganism increased the number and proportion of bacterivorous nematodes and decreased the number and proportion of plant-parasitic nematodes compared to traditional compost C. Total number of nematode was significantly influenced by compost amount, but didn't significantly influence by EM agent. Total abundance of nematode was positively correlated with the content of soil organic matter, total N, available P and K. The long-term addition of EM agent hasn't adverse effect on soil nematode community.

Keywords: effective microorganism; organic manure; soil nematodes; biodiversity

Introduction

The increasing demand of food supply following the increase of population in developing countries requires and needs a sustainable utilization of soil. However, the maintenance of costant levels of crop production or their increase is impossible without the use of soil fertilizers (Dik, 1992; Raun & Johnson, 1999). Long-term and large number of soil inorganic fertilizer applications can affect negatively soil fertility, soil biodiversity and crop products quality (Porazinska *et al.*, 1999; Gruzdeva *et al.*, 2007). So, considering the increasing importance of organic agriculture, soil organic manure applications are desirable and recommended. The effect deriving from organic manure applications includes numerous benefits resulting in an improvement of physical and chemical soil properties, i.e. porosity, aggregates stability, water exchange, fertility (Tester, 1990). Organic manure is an important source of nitrogen, phosphorus, calcium, and micro-elements such as zinc, copper, magnesium, which are essential to plant growth. Also, the incorporation of organic manure into the soil increases soil biological activity (Debosz *et al.*, 1999). Moreover, soil amendment with different types of organic manures caused a significant reduction of populations of plant parasitic nematodes, reduced nematode damages, improved plant yields and supported the beneficial free-living nematodes (Sasanelli, 1994; Akhtar, 1999; McSorley *et al.*, 1999; Sasanelli *et al.*, 2002; Arancon *et al.*, 2003; D'Addabbo *et al.*, 2003; Sasanelli *et al.*, 2006; Renčo *et al.*, 2007). Soil microorganisms play an important role in the cycle of nutrients and their activity strongly influences either plant growth or soil matter decomposition (Spedding *et al.*, 2004). Presence or addition of different useful microorganisms in the soil, called "effective microorganisms" (EM) as reported by Ni and Li (1998), could enhance nutrient elements availability. Therefore, the addition of EM to the traditional soil organic amendments could result in a positive synergic effect. Nematodes, abundant and ubiquitous in the soil are charac-

terised by several trophic behaviour and respond readily to environmental changes thus playing an important role in soil food web (Neher, 2001; Sánchez-Moreno *et al.*, 2006). In this system different groups of trophic nematodes are closely correlated with other soil organisms whose activities affect primary production, soil matter decomposition, energy flows, cycle of nutrients, especially nitrogen and phosphorus cycles (Sochová *et al.*, 2006). So, the variations of the nematofauna could provide important and useful information on soil health and quality (Urzelai *et al.*, 2000). Soil nematode community could be considered as a bioindicator of soil perturbations including chemical, physical and agricultural effects on ecosystem health and its functioning (Sánchez-Moreno *et al.*, 2009). The indicators include trophic group abundances and their ratios (Wasilewska, 1998), diversity indices (Urzelai *et al.*, 2000) and soil food web indices (Ferris *et al.*, 2001).

Some studies indicated that nematode abundance and diversity were influenced by application of organic and inorganic fertilizers (Forge *et al.*, 2005; Wang *et al.*, 2006; Okada & Harada, 2007; Liang *et al.*, 2008). Bulluck *et al.* (2002) and Forge *et al.* (2005) reported that populations of bacterivorous nematodes, mainly Rhabditidae and Cephalobidae, and fungivorous nematodes were greater in soils amended with manure than that in soils amended with chemical fertilizer. In some field experiments, Wang *et al.* (2006) and Okada and Harada (2007) observed that total nematode number and bacterivorous, fungivorous and omnivorous nematodes increased in conventional tillage system with chemical and organic fertilization of the plots in comparison to non amended plots. Few investigations have been undertaken on the effect of long-term application of organic manure on soil nematode populations (Liang *et al.* 2008). Therefore, a field experiment was conducted on wheat to assess the effect of long-term application of three different biological composts on soil nematode communities.

Materials and methods

The composts used were: traditional composted of 60 % straw, 30 % livestock dung, 5 % cottonseed-pressed trash and 5 % brans (C), traditional compost added with 200 ml aqueous solution of effective microorganism agents (EM) per 50 kg compost, according to the treatments (EMC) and chicken dung compost consisting of 60 % straw, 30 % EM

chicken dung, 5 % cottonseed-pressed trash and 5 % brans (EMCDC). Red sugar at dose of 1 kg was also added every 50 kg of both composts. Effective microorganism (EM) consisting of more 80 kinds of mixed and incubated microbes including photosynthetic microbes, acetate bacillus, actinomycetes, lactobacillus, microzyme (biological product for solids reduction and odor control of animal wastes), etc. as indicated by Ni and Li (1998). Effective microorganism agents were prepared as a commercial formulation in Beijing by Yiaimu Biotechnology Company.

The field experiment was set up in 1997 at the Qu-Zhou experimental station (36°52'N and 115°01'E), China Agricultural University, Hebei province, Northern China. The station is in a continental temperate monsoon zone and the climate in the region is warm, sub-humid and consists of summer rainfall and dry-cold winters.

Six treatments were arranged according to a randomized block design with three replicates per treatment. Plots (4 m x 8 m) were sown with winter wheat (*Triticum aestivum* L.) and summer maize (*Zea mays* L.) every year from 1997 to 2004. Treatments (Table 1 – 3) were: incorporation into the soil of compost added with EM (EMC), chicken dung compost added with EM (EMCDC) and traditional compost (C), each at the rates of 7.5 and 15 t/ha. Plots were treated every year, before sowing.

Soil samples, each consisting of fifteen cores (2.5 cm diameter x 20 cm deep), were collected from the upper soil layer (0 – 20cm) in June 2004 before sowing summer maize. Soil samples were put in plastic bags tightly closed to prevent moisture loss and stored in a refrigerator at a 4 °C until biological and chemical analyses. Then, root fragments and other organic debris were removed from each sample and the soil was thoroughly mixed before assessing soil moisture content and extracting free-living soil nematodes.

Soil moisture in each sample was gravimetrically determined by weight loss at 105 °C for 24 h and expressed as percent dry weight. Soil sub-samples were air-dried for 14 days at room temperature, and used to test alkaline hydrolysable N content, available P and K, organic matter, total N and soil pH. The potassium dichromate external heating method (Blakermore *et al.*, 1972), the semi-micro Kjeldahl method (Bremner, 1996), the alkaline hydrolysable diffusion method (ISSCAS, 1978), the classical Olsen method (Blakermore *et al.*, 1972) and the ammonium acetate flame photometry method (Lu, 1999) were used to determined

Table 1. Soil chemical properties in different fertilizer systems

Treatment	EMC	EMCDC	C	EMC	EMCDC	C
	(15 t/ha)			(7.5 t/ha)		
pH	7.16* **	7.16a	7.17a	7.18a	7.18a	7.20a
Organic matter (g/kg)	22.43a	22.50a	22.35a	19.31b	19.57b	19.23b
Total N (g/kg)	1.46a	1.46a	1.38ab	1.13c	1.20bc	1.09c
Alkali N (mg/kg)	129.52a	118.50ab	110.60abc	102.96bc	104.08bc	99.07c
Available P (mg/kg)	59.99a	60.48a	54.92a	31.00b	34.85b	24.66b
Available K (mg/kg)	256.02a	270.06a	278.08a	173.79b	181.81b	171.11b

EMC = Compost added with effective microorganisms; EMCDC = Chicken dung compost added with effective microorganisms; C = Traditional compost; * Average of three replicates; ** Data flanked in each row by the same letters are not significantly different according to LSD's Test (P=0.05).

organic matter, total N, alkaline hydrolysable N, available P and K, respectively. Soil pH was measured in 0.01 mol L⁻¹ CaCl₂ slurry (soil : solution = 1 : 2.5) using a glass electrode.

Nematode populations were extracted from 100 g composite fresh soil samples using the sugar flotation and centrifugation method (Barker *et al.*, 1985). The recovered nematodes were counted and preserved in a formalin aque-

ous solution (Steinberger & Sarig, 1993). One hundred randomly selected specimens per sample were identified, to genus level, using an inverted compound microscope (Mal & Lyon, 1975; Ying, 1998).

The characteristics of the nematode communities were described by the following approaches: 1) absolute abundance of individuals per 100 g dry soil; 2) trophic behaviour: a) bacterivores (BF); b) fungivores (FF); c) plant

Table 2. Mean relative abundance (%) of soil nematodes among all fertilizer systems

Treatment/genus	c-p value	EMC	EMCDC	C	EMC	EMCDC	C
		(15 t/ha)			(7.5 t/ha)		
Bacterivores		51.21*	56.06	46.07	59.56	58.49	39.48
<i>Cephalobus</i>	2	10.56	15.10	15.07	11.04	8.11	4.05
<i>Eucephalobus</i>	2	3.25	1.01	0.49	2.68	1.69	1.42
<i>Cervidellus</i>	2	1.28	0.00	0.00	0.50	0.00	0.00
<i>Acrobeles</i>	2	1.86	2.76	3.60	5.14	1.31	0.71
<i>Protorhabditis</i>	1	4.64	5.53	1.13	2.55	6.30	1.99
<i>Rhabditis</i>	1	29.61	30.49	23.81	36.56	40.44	30.01
<i>Plectus</i>	2	0.00	1.17	1.98	1.09	0.65	1.29
Plant-parasites		42.29	39.58	49.58	33.56	34.82	49.16
<i>Tylenchus</i>	2	1.97	1.01	2.38	1.36	2.40	2.34
<i>Tetylenchus</i>	2	0.00	1.01	0.00	0.00	0.00	1.23
<i>Belonolaimus</i>	2	1.28	0.58	1.49	0.00	1.25	0.00
<i>Tylenchorhynchus</i>	2	3.95	5.21	7.75	3.74	4.03	10.18
<i>Pratylenchus</i>	3	3.95	3.19	6.55	4.54	3.21	8.66
<i>Hoplolaimus</i>	3	8.92	13.21	8.01	10.40	4.38	6.86
<i>Helicotylenchus</i>	3	16.30	9.41	11.71	6.52	14.03	12.81
<i>Rotylenchus</i>	3	4.53	2.18	7.35	3.55	1.93	5.85
<i>Paratylenchus</i>	2	0.00	1.52	0.00	0.00	0.44	0.00
<i>Oxydirus</i>	5	0.00	0.00	1.75	0.00	0.65	0.00
<i>Longidorus</i>	5	0.69	1.67	2.11	3.45	2.50	0.64
<i>Xiphinema</i>	5	0.69	0.58	0.00	0.00	0.00	0.00
<i>Trichodorus</i>	4	0.00	0.00	0.49	0.00	0.00	0.58
Fungivores		5.23	2.76	3.85	3.15	3.97	6.97
<i>Ditylenchus</i>	2	0.69	0.00	0.49	0.00	0.00	0.00
<i>Aphelenchus</i>	2	1.28	0.58	0.63	1.19	0.00	2.52
<i>Aphelenchoides</i>	2	3.25	2.18	2.74	1.96	3.97	4.45
Omnivores-Predators		1.27	1.60	0.50	3.73	2.72	4.39
<i>Dorylaimus</i>	4	0.58	1.59	0.00	0.00	0.00	0.64
<i>Mesodorylaimus</i>	4	0.00	0.00	0.00	0.00	0.00	0.71
<i>Enchodelus</i>	4	0.00	0.00	0.00	0.00	0.65	0.00
<i>Eudorylaimus</i>	4	0.00	0.00	0.00	0.00	0.00	0.58
<i>Aporcelaimus</i>	5	0.69	0.00	0.50	3.73	2.07	2.45

EMC = Compost added with effective microorganisms; EMCDC = Chiken dung compost added with effective microorganisms;
C = Traditional compost; * Average of three replicates

parasites (PP); d) omnivores-predators (OP) (Yeates *et al.*, 1993; Pen-Mouratov *et al.*, 2004); 3) WI, ratio of bacterivores plus fungivores to plant parasites [(BF+FF)/PP] (Wasilewska, 1994); 4) fungivores/bacterivores ratio (FF/BF) (Twinn, 1974); 5) J', evenness, $J' = H'/\ln(S)$, where S is the number of taxa (Pielou, 1975); 6) λ , genus dominance, $\lambda = \sum pi^2$, where pi is the proportion of individuals in the i -th taxon (Simpson, 1949); 7) H' , Shannon index or diversity, $H' = -\sum pi(\ln pi)$, where pi is the proportion of individuals in the i -th taxon (Shannon & Weaver, 1949); 8) SR, species richness, $SR = (S-1)/\ln(N)$, where S is the number of taxa and N is the number of individuals identified (Yeates & King, 1997); 9) MI, maturity index, $MI = \sum vi pi$, where vi is the c-p value for free-living nematodes assigned by Bongers (Bongers, 1990) to the i -th nematode genus and pi is the proportion of the genus in the nematode community; 10) PPI, plant-parasitic nematodes maturity index, $PPI = \sum vi pi$, where vi is the c-p value for plant parasitic nematodes assigned by Bongers (1990) to the i -th nematode genus and pi is the proportion of the genus in the nematode community (Yeates & Bird, 1994); 11) $\sum MI$, modified maturity index, including plant-feeding nematodes, $\sum MI = \sum vi pi$, where vi is the c-p value for free-living and plant parasitic nematodes assigned by Bongers (1990) to the i -th nematode genus and pi is the proportion of the genus in the nematode community (Yeates, 1994). Enrichment index (EI), basal index (BI), structure index

(SI) and channel index (CI) were calculated as described by Ferris *et al.* (2001; 2004).

Data were also analyzed by two-way analysis of variance (ANOVA) and means compared by LSD's Test (Least Significant Difference). All statistical analyses were performed using SPSS 11.5 software package.

Results, discussion and conclusions

Soil chemical properties (Table 1) were not significantly impacted by EM agent, but soil organic matter, available P and K contents were significantly affected by the different amount of fertilizers.

Overall, 28 nematode genera were observed. In particular, 7 genera were bacterivores, 3 genera were fungivores, 13 genera were plant-parasites and 5 genera were omnivores-predators (Table 2). The bacterivores were the dominant trophic groups except in the traditional compost treatment. Among the different genera *Rhabditis*, *Cephalobus* and, *Helicotylenchus* were the most common. The number of identified taxa ranged from 17 to 21. Among treatments, bacterivores ranged between 39.48 % and 59.56 %, fungivores from 2.76 % and 6.97 %, plant-parasites from 33.56 % and 49.58 %, and omnivores-predators ranged between 0.50% and 4.39%. In particular, percentages of bacterivorous nematodes were 51.21 %, 56.06 %, 46.07 %, 59.56 %, 58.49 %, 39.48 % in EMC, EMCDC and C at

Table 3. Abundances of nematode individuals per 100 g dry soil, trophic groups and ecological indices for nematode community in different fertilizer systems

	EMC	EMCDC	C	EMC	EMCDC	C
	(15 t/ha)			(7.5 t/ha)		
Trophic groups						
BF	234.09a	302.37a	251.91a	269.79a	251.83a	145.87b
PP	193.65b	211.59ab	276.59a	149.62b	146.43b	183.34b
FF	23.85a	15.06a	21.79a	13.78a	16.91a	25.96a
OP	5.83ab	8.59ab	2.54b	16.98a	11.75ab	16.60a
TA	457.42b	537.61a	552.82a	450.17b	426.92bc	371.78c
Ecological indices						
WI	1.34a	1.67a	1.08a	1.91a	1.93a	0.96a
FF/BF	0.10ab	0.05b	0.10ab	0.06b	0.07b	0.18a
H'	2.24ab	2.18ab	2.26a	2.13ab	1.98b	2.19ab
λ	0.16b	0.16b	0.14b	0.18ab	0.22a	0.16b
J'	0.82ab	0.79ab	0.84a	0.82ab	0.77b	0.83a
SR	2.34a	2.33a	2.22a	2.02a	2.04a	2.20a
MI	1.47ab	1.45ab	1.53ab	1.58ab	1.40b	1.60a
PPI	2.90a	2.93a	2.91a	3.06a	2.94a	2.75a
$\sum MI$	2.07ab	2.02ab	2.23a	2.07ab	1.94b	2.16ab
EI	85.99ab	86.85ab	80.07b	86.89ab	92.05a	90.40a
BI	13.30ab	12.67ab	19.57a	11.72b	7.38b	8.50b
SI	18.67bc	23.33bc	7.94c	46.42ab	48.22ab	58.41a
CI	3.88ab	1.85b	4.29ab	2.07ab	2.21ab	5.14a

EMC = Compost added with effective microorganisms; EMCDC = Chicken dung compost added with effective microorganisms; C = Traditional compost; BF = Bacterivores; FF = Fungivores; PP = Plant parasites; OP = Omnivores-predators; TA = Total nematode individuals.

Community index: WI = ratio of bacterivores plus fungivores to plant parasites; FF/BF = fungivores/bacterivores ratio; H' = Shannon index; λ = genus dominance; J' = evenness; SR = species richness; MI = maturity index, PPI = plant-parasitic maturity index; $\sum MI$ = combined maturity index. including plant-parasitic nematode; EI = enrichment index; BI = basal index; SI = structural index; CI = channel index.

rates of 15 t/ha and in EMC, EMCDC and C at rates of 7.5 t/ha, respectively (Tab. 2). For the same treatments, percentages of plant-parasitic nematode were 42.29 %, 39.58 %, 49.58 %, 33.56 %, 34.82 %, and 49.16 % (Tab. 2).

The abundances of soil nematodes were significantly affected by the amount of composts, but they were not significantly influenced by EM agents (Tab. 3, Tab. 4). Bacterivorous, plant-parasitic and omnivorous and predatory nematode groups were significantly influenced by the amount of fertilization only in C treatment (Tab. 3). Significant EM agents effects were not found in the ratio of bacterivores plus fungivores to plant parasites, shannon index, genus dominance, evenness, species richness, maturity index, plant-parasitic nematode maturity index, modified maturity index, enrichment index, basal index, structure index and channel index. But significant compost amount effects were found in the fungivore/bacterivore ratio, genus dominance, enrichment index, basal index, and structure index (Tab. 3).

Total abundance of nematodes was positively correlated with the content of soil organic matter, total N, available P, and available K (Tab. 5).

The mean number of nematodes ranged between 372 and 553 individuals per 100 g dried soil, which is similar to results previously reported by Zolda (2006) in semi-natural steppe grasslands (185 – 590) and by Li *et al.* (2007) in rice-wheat rotation ecosystem (117 – 912). Abundances of

soil nematode were increased owing to addition of EM agents, except for EMC treatment at 15 t/ha.

The number of nematode genera (28) observed in this study was similar to that observed by Zhang *et al.* (2007) and Urzelai *et al.* (2000), but less than that reported for original dry meadow (56) (Zolda & Hánél, 2007), Japanese soybean field (51) (Okada & Harada, 2007) and long-term application of fertilizer (Liang *et al.*, 2008).

At the lower rate of compost the addition of EM agents significantly increased bacterivorous nematodes, whereas at 15 t/ha no significant increase was observed; in this case the positive effect of EM on BF could be masked by the larger amount of organic matter of the composts. Previous investigations reported that bacterivorous nematodes were more numerous than other trophic groups in soil treated with compost than in soil receiving chemical fertilizers (Ferris & Matute, 2003). Wang *et al.* (2006) found that bacterivorous nematodes were the predominant trophic group in a squash (*Cucurbita pepo*) agro-ecosystem and similar results were also reported by Jiang *et al.* (2007). However, Ou *et al.* (2005) found that plant parasitic nematodes were the dominant trophic group in a maize field.

In previous investigations the addition of EM agents also increased microorganism population in the soil (Tong *et al.*, 2003a; 2003b; Wang *et al.*, 2004; Zhou *et al.*, 2005).

Table 4. Two-way ANOVA analysis of soil nematode indices under EM effects and compost amount effects

Trophic groups	EM effects		Compost amount effects		Interaction	
	F	P	F	P	F	P
BF	4.307	0.039	3.290	0.095	3.446	0.066
FF	1.047	0.381	0.089	0.770	0.957	0.412
PP	3.317	0.071	11.224	0.006	0.501	0.618
OP	0.120	0.888	9.234	0.010	1.095	0.366
TA	1.228	0.327	42.816	0.000	10.981	0.002
Ecological indices						
WI	3.138	0.080	0.773	0.397	0.566	0.582
FF/BF	5.047	0.026	0.943	0.351	3.136	0.080
H'	1.526	0.257	3.484	0.087	0.321	0.732
λ	2.880	0.095	5.607	0.036	0.827	0.461
J'	3.390	0.068	0.363	0.558	0.377	0.694
SR	0.040	0.961	5.355	0.039	1.065	0.375
MI	2.654	0.111	0.696	0.420	0.828	0.460
PPI	0.981	0.403	0.002	0.962	1.029	0.387
$\sum MI$	3.223	0.076	0.417	0.531	0.142	0.869
EI	1.721	0.220	8.204	0.014	2.023	0.175
BI	1.578	0.246	10.371	0.007	2.213	0.152
SI	0.062	0.940	18.634	0.001	1.033	0.385
CI	3.272	0.073	0.054	0.821	0.881	0.439

BF = Bacterivores; FF = Fungivores; PP = Plant parasites; OP = Omnivores-predators, TA = Total nematode individuals.

Community index: WI = ratio of bacterivores plus fungivores to plant parasites; FF/BF = fungivores/bacterivores ratio; H' = Shannon index; λ = genus dominance; J' = evenness; SR = species richness; MI = maturity index, PPI = plant-parasitic maturity index; $\sum MI$ = combined maturity index, including plant-parasitic nematode; EI = enrichment index; BI = basal index; SI = structural index; CI = channel index.

Values of $P < 0.05$ were considered significantly.

Table 5. Correlation coefficients between nematode indices and soil chemical properties

	pH	Organic matter	Total N	Alkali N	Available P	Available K
Trophic groups						
BF	-0.287	0.398	0.352	0.390	0.482*	0.391
FF	-0.137	-0.022	-0.007	-0.109	-0.068	0.027
PP	-0.136	0.367	0.428	0.174	0.491*	0.554*
OP	-0.210	-0.511*	-0.395	-0.573*	-0.592**	-0.564*
TA	-0.415	0.605**	0.628**	0.420	0.768**	0.753**
Ecological indices						
WI	-0.172	0.020	-0.074	0.081	-0.035	-0.104
F/B	0.071	-0.237	-0.220	-0.323	-0.337	-0.198
H'	-0.036	0.230	0.217	0.052	0.184	0.297
λ	0.045	-0.301	-0.210	-0.069	-0.189	-0.365
J'	0.108	-0.013	-0.130	-0.156	-0.149	0.009
SR	-0.160	0.374	0.517*	0.243	0.422	0.437
MI	-0.149	-0.039	-0.165	-0.428	-0.363	-0.062
PPI	-0.218	-0.007	0.158	0.183	0.117	0.047
$\sum MI$	-0.097	0.014	0.118	-0.166	-0.009	0.154
EI	0.012	-0.598**	-0.293	-0.179	-0.366	-0.610**
BI	0.003	0.622**	0.320	0.221	0.419	0.643***
SI	-0.133	-0.550*	-0.431	-0.585*	-0.684**	-0.623**
CI	0.025	0.011	-0.092	-0.183	-0.148	0.035

BF = Bacterivores; FF = Fungivores; PP = Plant parasites; OP = Omnivores-predators, TA = Total nematode individuals.

Community index: WI = ratio of bacterivores plus fungivores to plant parasites; FF/BF = fungivores/bacterivores ratio; H' = Shannon index; λ = genus dominance; J' = evenness; SR = species richness; MI = maturity index, PPI = plant-parasitic maturity index; $\sum MI$ = combined maturity index, including plant-parasitic nematode; EI = enrichment index; BI = basal index; SI = structural index; CI = channel index.* Significant at $P < 0.05$ and ** significant at $P < 0.01$.

The decreasing effect of effective microorganism on number of plant-parasitic nematodes in comparison to traditional compost treatments could be attributed to some toxic effect of metabolites produced by EM and/or to a competition between the two groups, indicating that EM agents could play a role in the biological control of plant parasitic nematodes.

Mean WI values ranged between 0.93 and 1.67, and are comparable to the mean values observed by Liang *et al.* (1999) for a potato field and by Pen-Mouratov *et al.* (2008) in a desert communities. The FF/BF values reflect the structure of the microflora community and indicate the status of the decomposition pathway in detrital food webs (Ruess, 2003). They ranged between 0.05 and 0.18 and were lower than that observed in desert ecosystem (Pen-Mouratov *et al.*, 2008), indicating a large bacterial population presence with a dominant bacterial decomposition pathway (Ruess, 2003). The T values (2.06 – 2.45) observed in our study were similar to those (2.01 – 2.37) reported by Pen-Mouratov *et al.* (2008) in the Northern Negev Desert, Israel.

The Shannon index (H') affords more weight to rare species, and its high value indicates greater diversity. The H' values (1.98 – 2.26) observed in this study were higher than that observed by Bulluck *et al.* (2002) in Goldsboro and Clinton farming systems (0.55 – 1.47). Genus dominance (λ) was used to assess the dominance of all nematode genera in the samples. The observed λ values (0.14 – 0.22) were within the range (0.09 – 0.70) observed by

McSorley & Frederick (1996) and lower than that obtained by Pen-Mouratov *et al.* (2004). The J' values (0.77 – 0.84) calculated in the field experiment were higher than results reported by Pen-Mouratov *et al.* (2008), whereas the species richness (SR) values (2.02 – 2.34) were lower than that reported by Yeates & King (1997) (3.03).

The maturity index (MI) is a measure based on the composition of the nematode community and can reflect the degree of disturbance of the soil ecosystem (Bongers, 1990). In this study the MI values ranged from 1.40 and 1.58 and they were lower than that reported in Chinese paddy fields (2.86 – 3.23) (Liu *et al.*, 2008). The lower MI values observed suggested that abundant opportunist nematodes (c – p values 1 – 2) were present in compost treated soil. The plant parasitic indices (PPI) ranged from 2.75 to 3.06 and $\sum MI$ values ranged from 1.94 – 2.23. At 15 t/ha of the composts, plant parasitic populations were significantly decreased by the addition of EM to the traditional organic compost C. PPI and $\sum MI$ values were higher and similar to that observed in the black soil region of Northeast China by Liang *et al.* (2002) and by Wang *et al.* (2006), respectively.

In conclusion, the application of EM to organic compost could increase the number and proportion of bacterivorous nematodes, and decrease the number and proportion of plant-parasitic nematode compared to traditional compost alone. Therefore, long-term addition of EM agents could be suggested as a nematode management strategy as it provides an increase of pest suppressivity by reducing the

impact on soil-plant equilibrium compared to chemicals and inorganic fertilizers.

Acknowledgements

This research was supported by the Key Project of the National “10th-FiveYear-Plan” of China (Grant no. 2004BA508B01) and Key Discipline Foundation of Ecology in Beijing Municipality (Grant no. XK10019440). The authors wish to express their appreciation to anonymous reviewer for valuable comments.

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RECEIVED FEBRUARY 27, 2009

ACCEPTED OCTOBER 30, 2009